

 The Power of Commitment

Waste Futures – Green Island Landfill Closure

Groundwater Technical Assessment – October 2024 Update

Dunedin City Council

18 July 2024

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1. Introduction

1.1 Background

As part of Dunedin's wider commitment to reducing carbon emissions and reducing waste going to landfill, the Dunedin City Council (Council) has embarked on the Waste Futures Programme to develop an improved comprehensive waste management and diverted material system for Ōtepoti Dunedin. The Waste Futures Programme includes the roll out of an enhanced kerbside recycling and waste collection service for the city from July 2024. The new service will include collection of food and green waste.

To support the implementation of the new kerbside collection service, the DCC are planning to make changes to the use of Green Island landfill site (Figure 1.1) in coming years.

Figure 1.1 Site location

The proposed changes include:

- planning for the closure of the Green Island landfill, which is coming to the end of its operational life
- developing an improved Resource Recovery Park (RRPP) to process recycling, and food and green waste
- providing new waste transfer facilities to service a new Class 1 landfill currently planned for a site south of Dunedin, at Smooth Hill.

The resource consents for the new Smooth Hill landfill were granted in May 2023. Depending on DCC decisions regarding development of Smooth Hill, time needed to undertake baseline monitoring, preparation of management plans, landfill and supporting infrastructure design and construction, DCC anticipate that the new Class I landfill facility, won't be able to accept waste until 2027/2028 at the earliest.

In the interim, DCC therefore plans to continue to use Green Island landfill for waste disposal. Based on Dunedin's current waste disposal rates, it is likely that that the Green Island landfill can keep accepting waste for another six years (until about 2029). Between now and then, and as it continues to fill up, the landfill will be closed and capped in stages. When the landfill closes completely, there will be opportunities for environmental enhancements and public recreational use around the edge of the site. Examples could be planting restoration projects and new walking and biking tracks beside the Kaikorai Estuary. Long term use and public access to the landfill site post closure will be determined in consultation with Te Rūnanga o Ōtākou, the local community and key stakeholders.

As current Otago Regional Council resource consents needed to operate a landfill at Green Island expire in October 2023, the DCC are now applying to ORC for replacement resource consents to continue to use the landfill until it closes completely, and waste disposal can be transferred to a new landfill facility. The replacement consents relate to ground disturbance, flood defence and discharges to land, water, and air. The site is subject to an operative designation (D658) in the Proposed Second-Generation Dunedin City District Plan (2GP) for the purpose of Landfilling and Associated Refuse Processing Operations and Activities.

The development of the new RRPP and waste transfer facilities at Green Island does not form part of the replacement consent applications. Resource consents for the development and operation of the RRPP were submitted in March 2024 and are under consideration by ORC..

This groundwater technical assessment has been prepared to support the resource consent application for the continued operation and closure of Green Island Landfill. This report has been updated in October 2024 to include new information and to respond to s92 questions from the ORC. This version of the report replaces the original Groundwater report issued in March 2023.

1.2 Purpose of this report

The purpose of this report is as follows:

- Provide a description of the groundwater environment, existing management of groundwater and leachate discharges, and environmental monitoring data.
- Provide a technical assessment of the potential effects on groundwater and connected surface water flows from the proposed extension of landfill operations and closure management.
- Provide an assessment of the effects of leachate leakage on groundwater quality.
- Provide recommendations for mitigation measures to minimise the effects on the environment and monitoring conditions to confirm the effectiveness of the recommended mitigation measures.

Note: the potential effects of site stormwater on downstream surface water flows and quality are discussed in the *Surface Water Technical Report* (GHD 2024) and referenced in this report. This Surface Water Report has also been updated in October 2024 with additional information and in response to s92 questions from the ORC.

This report should be read in conjunction with the following reports:

- Green Island Landfill Design Report (GHD, 2023)
- Surface Water Report (GHD, 202), including Appendix B, Annual Monitoring Plan 2022-2023.
- Green Island Landfill Geotechnical Factual Report (GHD, 2023)
- Green Island Landfill Liquification and Stability Assessment (GHD, 2023)
- Human Health & Environment Risk Assessment (GHD, 2024)

These reports provide supporting information and context which the surface water assessment relies upon. Where appropriate, a summary of critical information is summarised in this report with crossed references to the relevant technical report.

In addition, the assessment undertaken in this report is based on a review of previous investigations, including those undertaken as part of the 1994 resource consent application. This information has been supplemented with additional site investigations undertaken by GHD to support the design and consenting process (documented in the *Geotechnical Factual Report* (GHD 2023). The information obtained from the desktop review and site investigations has been used to undertake an assessment of potential impacts to groundwater and connected surface water, associated with the proposed extension and ultimate closure of Green Island Landfill.

1.3 Current landfill operation and management

1.3.1 Current Consents

The operation of the Green Island Landfill, including associated waste processing operations and facilities, is currently subject to 14 existing resource consents granted by Otago Regional Council (ORC). The consents cover landfill operation activities relating to discharges to land, water, and air, taking and/or diverting water, and disturbance of a contaminated site. All consents expire on 1 October 2023.

The current consents limit the extent of landfilling through the combination of a maximum 38 ha landfill footprint, conditions limiting the deposit of waste to 270 m³/day and 100,000 m³/year¹, and the 2023 term of the consents. The consent conditions do not impose any specific limit on the overall finished height, shape, or contour of the landfill. However, the plans included in the 1994 resource consent applications show a finished landfill surface rising to a maximum height of 25 m above mean sea level (amsl).

The consent conditions also require the consents are exercised in accordance with a Landfill Work Programme (LWP) prepared by the consent holder, which is to be reviewed annually or at such lesser frequency as the consent authority may approve. Among other matters, the LWP is required to describe present projections and intentions for landfill operations, and the sequencing of works.

1.3.2 Landfill Development and Management Plan

A Landfill Development and Management Plan (LDMP) was developed following the issuing of the consents to serve the purpose of the LWP. The LDMP is to document site-specific procedures, including monitoring and contingency actions to be implemented to ensure the landfill achieves the conditions set out in the resource consents. The LDMP is structured into the sections set out below:

- 1. **Introduction** the existing resource consents, designation, and status and review of the LDMP.
- 2. **Site Management** management structure, responsibilities, requirements for staff training, and community liaison.
- 3. **Landfill Development** including design principles, landfill capacity, and the filling programme and sequence.
- 4. **Site Operations** including controls and procedures for access control, stormwater management, leachate management, LFG management, greenwaste mulching and composting, salvage and management of diverted materials, roading and traffic management, waste acceptance and placement, waste cover, and control of nuisances.
- 5. **Environmental Monitoring** including monitoring, recording, and reporting for surface water, groundwater, LFG, leachate, odour, and weather.
- 6. **Emergency Management** including procedures for management of fires, hazardous waste/materials, leachate and LFG escape, extreme weather/flooding, machinery failure, accidents, and earthquakes.
- 7. **Closure, Reinstatement, and Aftercare** including final capping, continued operation and maintenance of landfill infrastructure, and ongoing monitoring.

The LDMP was first provided to ORC in 1994 following the issuing of the consents and was subsequently updated in 2004, and 2007. The most recent LDMP, which reflects the current approach to landfill operation and management was provided to ORC in February 2023.

1.3.3 Landfill Operations Plan

The landfill is currently operated by Waste Management NZ Ltd. under contract to the Council. Waste Management NZ Ltd. are required to maintain a Landfill Operations Plan (LOP) which reflects the LDMP and more specifically addresses day-to-day management landfill operational matters.

The LDMP (February 2023) and LOP (October 2018) will be updated after the granting of any replacement resource consents to ensure that they align with the final approved consent documentation, and any resource consent conditions.

2. Environmental Setting

2.1 Introduction

A review of the environmental setting was undertaken to inform the technical assessment (Section 3). This review includes a factual summary of information relevant to the conceptual understanding of the groundwater system and baseline parameters used in the modelling assessment. Assessment and interpretation of the site data is included in the following sections.

2.2 Site Description

The Green Island Landfill site is located in the suburb of Green Island, approximately 8.8 km by road southwest of Central Dunedin. The landfill site comprises a total area of 75.6 Ha. The site is generally bound by State Highway 1 to the north, the Kaikorai Stream and Estuary to the west, the Green Island Wastewater Treatment Plant (GIWWTP) to the southwest, Brighton Road to the south, and the Clariton Ave residential area and Brighton Road industrial area to the east.

The margins of the Kaikorai Stream and Estuary bordering the landfill to the north and west are identified as a Regionally Significant Wetland in the Regional Plan: Water; and an Area of Significant Biodiversity Value, and a Wāhi Tupuna of cultural significance to mana whenua in the 2GP. Low lying areas around the stream and estuary are also identified as being within a Hazard 2 Flood overlay at risk of flooding in the 2GP.

2.2.1 History

The historical placement of waste and its distribution across the site is described in detail in Appendix D. The following provides a summary of the waste filling history that is relevant to the future engineering design and closure management presented herein.

Waste disposal first occurred at the Green Island site in 1954 with the disposal of industrial waste and the site has been used for waste disposal since that time. A number of other sites have been used over the decades across the Dunedin region including the "Maxwell" landfill on the opposite side of the estuary to Green Island landfill. The Maxwell landfill was formally closed to waste disposal in mid-2017 and the Green Island landfill continued as the sole municipal solid waste disposal facility in the Dunedin region after that time. The existing operational consents were granted in 1994.

Landfilling commenced at the south-east corner of the landfill site and has continued north and west over the decades. The eastern portion of the landfill has a relatively shallow depth of waste at around 3 to 6 m thickness and is currently used for facilities and waste transfer station operations. This area is proposed to be developed in the near future to establish the Resource Recovery Park Precinct (RRPP). No further waste disposal will occur in this area.

The main landfill area is located immediately to the west of the facilities area. Waste placement in this area has been confined over recent decades within a constructed soil bund that encircles the landfill on the eastern, northern and western sides adjacent to the estuary. However, prior to berm construction waste had been placed across the whole extent of the area (see Appendix D). In recent years significant waste disposal has progressed north to south. In the northern and eastern areas waste has been placed up to the 1994 design contours (see *Design Report (GHD 2023))* and final capping has been completed. The south western half of the landfill has up to approximately 6 m - 8 m depth of waste placed during the 1990's, and a further 10 to 15 m of waste can be placed in this area. This is the primary area where future waste placement will occur through to closure of the landfill.

The pre-existing landform for the Green Island landfill was tidal estuary associated with the upper reaches of the Kaikorai Estuary (Figure 2.2). Abbotts Creek flows into Kaikorai Stream to the north of the site with the Kaikorai Stream flowing to the east then south in the Kaikorai Estuary. The 1942 aerial photograph shows a very similar Kaikorai Stream alignment to present. Several small channels are evident on the tidal mudflats, these drain towards the estuary at the southwest corner of the site. The 1942 photo also shows some evidence of land disturbance (to the south) and drains or trenches that predate the landfill activities.

Waste was originally end dumped directly onto the estuarine muds and up against the south eastern estuary edge where the pre-existing landform rises gently to the southeast. As discussed above, a soil bund was constructed around the north, west and south-western sides of the landfill to confine the waste from the adjacent Kaikorai Stream. The current landfill has an access track on the outside of the bund along with a leachate trench which is part of the leachate collection system and perimeter groundwater monitoring wells.

The leachate trench was installed on the outside of the soil bund in the mid-1990's. This perimeter control is not present along the southern side of the landfill against the rising ground of the hillside *(see Drawing 12547621-01- G102, Appendix A Design Report (GHD 2023*). The main wastewater trunk sewer follows the existing southern extent of existing landfill, flowing to the GIWWTP located 200 m southwest of the landfill site. A surface water drain follows the alignment of the sewer creating a valley that intercepts runoff from the landfill and directs it to the leachate collection system.

Figure 2.2 1942 Aerial photo showing pre existing landform.

2.2.2 Leachate management

The perimeter bund and leachate trench was installed on the outside of the soil bund in 1994 and commissioned in 1995. The leachate trench intersects contaminated groundwater (landfill leachate mixed with groundwater) seeping from the site. The leachate collection system comprises the gravel interception trench with the HDPE liner (on outer side) and the slotted PVC drainage pipe, together with a manhole and pump station configuration (Figure 2.3). Figure 2.4 shows details of the gravity leachate trench drain arrangement between pump stations and manholes. The manholes are located at the ends of each PVC collector pipe, allowing inspection and cleaning of the pipes to occur. Impacted groundwater is then conveyed by gravity to nine individual pump stations, which then pump into a 125 mm dia. rising main, which conveys the leachate and groundwater to the main sewer and ultimately GIWWTP. The rising main has a discharge into the sewer at each end, one approximately 79 m south of PS1 and the other approximately 80 m south of PS8. Hence, the riser pumps in both clockwise and/or

anti-clockwise direction before being discharged to the sewer line. PS9 discharges directly to the sewer main in the southern valley.

Figure 2.3 Leachate drain schematic (City Consultants, 1997)

The trench creates a hydraulic barrier for groundwater and leachate migration offsite. The HDPE liner aids in reducing the volume of water entering the trench from the Kaikorai Stream but does not completely prevent inflows. The continuous dewatering of the trench is required to maintain this barrier, with the pump stations set to maintain water levels at low levels to create the hydraulic gradient which directs flow to the trench. The leachate pumps are automated to start when the pumpwell reaches a "Pump ON" level, and then stop when the level reaches a "Pump OFF" level. The pumpstations also have alarms for the following conditions (high level alarm, pump running low level and power loss). Flow rate through the pump is measured continuously, as are the pump run hours. Consent conditions require regular monitoring of groundwater levels adjacent to the trench to confirm the hydraulic gradient (discussed in sections 2.5.3).

The trench is installed in the Upper Kaikorai Estuary Member (UKEM) (see section 2.4) comprising fine sands and silt. However, landfill refuse was recorded as overlying the UKEM in over half of the trench profiles during construction (Barry Douglas, 2022) with a maximum thickness of landfill recorded of 2.6 m. Borehole data from the recent drilling investigation (see section 2.4.2) confirmed similar profiles.

The leachate trench is absent along the southern edge of the landfill where waste is placed against the base of the slope that rises to the east. Management of leachate in this area is currently via a shallow surface drain which conveys the leachate (and any shallow groundwater seepage) to PS1. It is also noted that there is a 90 m gap in the trench between MH8 and PS9. This gap aligns with a short ridge of land that extended into the estuary based on historical maps and photos. Based on the geological map this ridge is inferred to be mudstone.

A culvert located on the eastern side of the landfill between the South Eastern Constructed wetland and the Eastern Constructed Wetland has recently been identified as a pathway for leachate seepage, which has been confirmed from water quality monitoring and a culvert inspection (discussed in the *Surface Water Report, 2024*). These results and proposed mitigation are discussed further in the *Design Report* (GHD 2023). As at October 2024 work is underway to repair this culvert and work is expected to be complete by end of March 2025.

Additional perforated leachate drains have been installed over intermediate cover soils in the southern portion of the landfill and in the northern sector of waste placed in 2019-2022. These drains direct leachate to the perimeter leachate collection trench. These drains installed within the landfill are described in the *Design Report (GHD 2023)* and shown on Drawings 12547621-01-C204 (*Design Report – Appendix A).*

Figure 2.4 Well and pump station arrangement, leachate collection system

NTS

Figure 2.5 Typical cross section of leachate collection system (MWH, 2004

2.2.3 Leachate volumes

Pump flow rates and pump hours are recorded within each pump station continuously. The volume of leachate pumped from each pump station to the rising main (weekly total) was reviewed as part of this assessment. Based on this record, a box and whisker plot of the combined flow rate (to all pump stations) averaged on a weekly basis is presented in Figure 2.6. Lower flow rates in the 2020-2021 and 2021-2022 reporting years are likely to be related to below average rainfall during this period (Table 2.1) but may also partly reflect the installation of final capping to part of the landfill which will reduce rainfall seepage into the waste. It is understood that PS1 has in the past received large quantities of stormwater flows from the landfill via open drains, with a recent improvement to the stormwater collection from the landfill (September 2021) resulting in some of these flows now going directly to PS3. Water held in the Northern Leachate Pond (see Drawing 12547621-01-C402) is also diverted to the leachate collection system (at PS5). Table 2.2 lists the median flow rates from the individual pump stations for an average rainfall year (2019-2020) and the past year (2021-2022, below average rainfall).

Table 2.1 Rainfall recorded at Musselburgh EWS

1 Average rainfall ~750 mm

Table 2.2 Pump station flows – Median flow rate (reporting year July to June)

2.3 Hydrology

As discussed above the Kaikorai Stream borders the landfill to the north and west. Summary flow statistics for the Kaikorai Stream are provided below in Table 2.3 (NIWA, 2023¹). Further details on the surface water environment

¹ NIWA River Maps online view: **https://shiny.niwa.co.nz/nzrivermaps/** (accessed 09/02/2023)

are provided in the *Surface Water Technical Assessment (GHD, 2023)*. From time to time, the estuary/ river mouth is blocked as a result of sand/debris accumulation from storm events. When this occurs, water backs up in the stream and results in higher stream levels adjacent to the site. Elevated water levels are maintained until the mouth is opened up again, either by natural or mechanical methods.

2.4 Geology

2.4.1 Overview

The site is situated on estuarine sedimentary deposits associated with the Kaikorai Estuary (Figure 2.7). These have been characterised as the Kaikorai Estuary Formation (KEF) (BDGC, 2002). KEF was considered to extend to a depth of approximately 11 m in the landfill area, and immediately overlies the Abbotsford Formation mudstone (BDGC, 2002)

The KEF was divided into an upper and lower layer (Member), with the upper member being further divided by BDGC (2020) into two subgroups as shown in Table 2.4. The estuarine sediments are underlain by weak mudstone rocks belonging to the Abbotsford Formation(Adams Geotechnical, 2019; Geotechnical Report (GHD 2023)).

To the southeast of the site, the landfill and underlying alluvium terminate at the toe of a 30m high ridge that has been mapped as the upper part of the Abbotsford Formation sequence (Adams Geotechnical, 2019) as shown in Figure 2.8. The Abbotsford Formation is described as "Grey to dark grey sandstone, siltstone and claystone with some glauconitic mudstone and green sand layers" (McKellar, 1990). Overlying the mudstone is a layer of loess and colluvium (clay -silt soils). The thickness of the colluvium/loess layer is variable, near the base of the ridge (southern valley) the Test pit investigations by Adams Geotechnical indicated that the colluvial layer was generally 1-2 meters thick. Groundwater was not intercepted in any of the test pits (Adams Geotechnical, 2019).

Member	Description	Subgroup	Thickness
Upper Kaikorai Estuary Member (UKEM)	Variable thin beds of sand, silty sand, sandy silt, silt, clayey silt and silty clay	Subgroup A -mostly homogeneous fine grained	4.5 _m
		Subgroup B – heterogeneous, coarser grain size	
Lower Kaikorai Estuary Member (LKEM)	Massive homogeneous beds of clayey silt, silty clay and silt, and minor (possibly localised) beds of clay, very fine sandy silt and silty very fine sand.	$\,$	6.5 m

Table 2.4 Description of KEF lithological units (after BDGC, 2002)

Figure 2.7 Geological map of Green Island area (McKellar, 1990). Approximate site boundary shown by red dashed line. ab: Abbotsford Formation f: alluvium, xd:fill

Figure 2.8 Interpreted geological cross section from Adams Geotechnical (2019)

2.4.2 GHD site investigation (2022)

Site investigations were undertaken by GHD to inform the geotechnical and hydrogeological assessments. The results of the site investigation are documented in the *Geotechnical Factual Report* (GHD 2023). A brief summary of the geology encountered is included here. The GHD investigation comprised twelve bore holes across the site, with piezometers installed in six of the bore holes. The general geological profile for boreholes on site perimeter is summarised in Table 2.5 below. There was no clear geological distinction between the two subgroups of the Upper Kaikorai Estuary Member (UKEM), therefore we have not adopted this division. In several of the bore holes there was a coarse grained layer (sand and/or gravel) at the contact of the Lower Kaikorai Estuary Member (LKEM) and mudstone. Depending on the location, variable amounts of fill were encountered. At BH104, drilled in the south-eastern part of the site, fill sits directly on mudstone.

Bore locations are included in Appendix B-1.

Table 2.5 Summary geological profile – perimeter (near leachate trench) boreholes (Geotechnical Factual Report, GHD 2023)

2.5 Hydrogeology

2.5.1 Overview

Prior to the landfill, groundwater within the estuarine deposits (KEF) is likely to have been hydraulically connected to the Kaikorai Stream and other surface water features. As discussed in Section 2.2.1, pumping from the perimeter leachate trench creates a hydraulic barrier between surface water and the shallow aquifer underlying the landfill. The underlying Abbotsford Formation is inferred to be an aquitard due to the very low permeability of the mudstone/siltstone and is effectively an impermeable barrier for downward seepage. Adams Geotechnical (2019) reported permeabilities for the Abbortsford Formation siltstone associated with the capping material borrow pit located to the east of the landfill, between 3.8-7.9 x 10-10 m/s.

2.5.2 Hydraulic Conductivity

Groundwater investigations undertaken as part of the 1992 EIA (Beca, 1992), included permeability testing in three boreholes at various depths. The results indicated an average hydraulic conductivity for the estuarine sediments of 4×10^{-6} m/s above 4 m depth and 6 x 10⁻⁷ m/s for greater than 4 m depth.

Hydraulic conductivity estimates of the differing geological units at the site, determined from hydraulic testing during the GHD site investigation (Appendix A) are presented in Table 2.6. The testing resulted in a range of hydraulic conductivity values; this reflects the heterogeneity created by the depositional environment. There are likely to be discrete channels of higher permeability materials (where active stream channels were located), with both horizontal (due to the direction of deposition) and vertical anisotropy² (due to the layering of sediments) likely.

While investigations beneath the landfill waste are necessarily limited it is expected that the permeability of the estuarine sediments is likely to be reduced beneath the landfill footprint due to the compression of the sediments from the weight of the landfill.

 2 Where the hydraulic conductivity of sediments is different in two directions. Ie. in the horizontal direction compared to the vertical direction, in layered sedimentary units the vertical to horizontal hydraulic conductivity ratio is commonly 0.1.

2.5.3 Groundwater monitoring

2.5.3.1 Monitoring network

Monitoring of water levels and water quality is undertaken on a routine basis in accordance with the conditions of the current consents. The monitoring programme includes sampling of surface water, groundwater and leachate. A review of the surface water quality data is presented in the *Surface Water Technical Assessment* (GHD, 2023), and surface water quality will only be discussed here where relevant to the groundwater assessment. The groundwater monitoring network comprises:

- Eight lines of groundwater monitoring wells transecting the leachate collection trench, as shown in **Error! Reference source not found.**.
- Each Well Line is located at mid-distance between two pump stations and each line comprises three shallow wells, MWA through to MWC, with the exception of Line 7, where MWC is absent (noting that the majority of these are founded in the UKEM geological unit).
- At each Well Line, monitoring wells MWA and MWB are located on the landfill side of the leachate trench, approximately 20 m and 5 m from the trench respectively.
- Monitoring well MWC is located between the trench and the Kaikorai Stream / eastern sedimentation pond / eastern boundary.
- Along each Well Line, an inspection manhole is located at the point the Well Line intersects the leachate trench, between monitoring wells MWB and MWC.
- On three of the Well Lines (Well Line 2, 4 and 7), deep wells are also present and monitored, located between the leachate collection trench and the stream. They are described as MWD (and founded in the LKEM geological unit).
- An additional bore, MW0C located at the end of the leachate trench collection system at Well Line 0, to the south of PS1 is also monitored.
- A further monitoring well, MW9D, had been located towards the centre of the landfill, but was lost due to landfilling activities in 2015.

Bore locations are included in Appendix B. A schematic of the monitoring Wells Lines is shown in Figure 2.9.

Figure 2.9 Schematic of monitoring well transect

2.5.3.2 Groundwater levels

The leachate collection system was designed to create a hydraulic barrier by reducing groundwater levels in the trench and drawing groundwater from both sides of the trench (i.e. landfill and stream side, although the stream side inflows were designed to be retarded by the installation of a HDPE liner on the stream side of the trench). Groundwater levels are checked on a monthly basis to monitor the hydraulic gradient. The water level monitoring consistently shows that the lowest groundwater levels (in monitoring wells) occur adjacent to the trench (MWC wells). These results are shown in the Green Island annual monitoring report (GHD, 2022 and attached to the Surface Water Report (GHD 2023)) with a selection representing winter, spring, summer, and autumn, included in Appendix B. These groundwater levels have been monitored monthly and reported annually to the ORC since the commencement of these consents in 1994.

Long term groundwater level records were reviewed as part of this assessment. Groundwater level data is available for the period from 1995 to 2003 and from 2015 to present. In general:

- Groundwater levels fluctuate within a range for each well, with no long-term trend in groundwater levels evident (with the exception of MW4D discussed below, Figure 2.11); and
- Seasonal variation is evident in the record for some wells (e.g. MW3C and MW6C, Figure 2.12). Groundwater levels are generally lowest in drier periods (summer-autumn), with groundwater highs occurring winter/spring and large rainfall events (e.g. January 2021).

Groundwater levels in MW4D showed an increasing trend from 1997 to 2003 (Figure 2.11). Due to the increasing groundwater head, the monitoring well casing was extended higher above ground to ensure no leakage of groundwater from the top of the casing (artesian conditions). The increasing artesian conditions are likely to be a function of the weight of the landfill squeezing and compressing the underlying estuarine sediments. Observed changes in groundwater level (shown in Figure 2.11) appear to correlate well with wate placement in this sector of the landfill. There is some uncertainty about the datum used for 2015-2016 measurements³, however the data shows that groundwater levels appear to have stabilised around 1.7- 2.2 m RL (NZVD⁴) in recent years. MW4D was drilled to a depth of 10.5 m depth. There is no bore log available for the well, but the geology encountered in a nearby borehole, BH100 (*Geotechnical Factual Report* GHD,2023), suggests that MW4D is screened over a sand and gravel layer present between the base of the LKEM and mudstone.

 3 Note throughout this report two datums are used. On older figures/drawings a DCC Design Datum of AMSL +100m is used (hence a 1994 flood level of 103.3m). More recent data and the design drawings for this study use NZVD2016 as the datum and are referred to through this report as "amsl"

⁴ Elevation of MW4D converted to NZVD based on correction factor of -(0.388 m + 100 m), from height of mark AG1B **AG1B: W 6 (linz.govt.nz)**

1,399,000

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Legend

Monitoring Well **Allected River** Site Boundary \bullet

Figure 2.11 Groundwater levels measured in MW4D

Figure 2.12 Groundwater levels measured in MW3C and MW6C – 2015 to 2022

2.5.3.3 Groundwater quality

Groundwater quality is monitored on a quarterly basis in accordance with the consenting conditions, and results are reported in annual monitoring report, the most recent being the 2022-2023 report (GHD, 2023). The annual report provides a full summary of water quality trends and comparison against relevant standards/guidelines. A brief overview of the data is provided below with selected monitoring plots included in Appendix C.

In general, the monitoring wells inside the leachate trench (A and B series) are likely to be impacted by leachate. However, some of the wells on the outside of the trench also show the influence of landfill waste. A review of the site history and waste distribution was undertaken (included as Appendix D) which identified areas where waste has been placed outside of the leachate trench. The likely extent of waste on the outside of the trench is shown on Figure 2.13. Based on this review, the following "C" wells are likely to be within or influenced by historical waste materials; 8C, 7C, 6C, 4C, and 3C. Of the "D" wells, 7D is only 7 m deep and may be within waste. Monitoring well 4D, at 10.5 deep, is likely to be screened below the waste materials.

The impact of waste outside of the trench is managed through operation of the leachate collection system, which pulls groundwater (and any leachate) from both sides of the trench. This is discussed further in Section 4.

Water quality trends

A review of water quality data was undertaken to inform the conceptual understanding of the groundwater system and interactions with leachate/water. The water quality trends and patterns for relevant parameters are summarised below in Table 2.7. Water quality plots are provided in Appendix C.

Parameter	Groundwater trends	
Electrical Conductivity	Elevated in all monitoring wells relative to typical background groundwater Deep wells - highest in 2D and 4D (also higher chloride in these wells) Shallow wells - no clear pattern between A/B/C wells	
Dissolved oxygen	Dissolved oxygen in groundwater is low with many samples <20% oxygen saturation, in contrast most surface samples are > 50%. The Eastern Constructed Wetland and South Eastern Constructed Wetland exhibit a wide variation in dissolved oxygen content.	
Ammoniacal nitrogen	Generally elevated in groundwater relative to surface water with the exception of Eastern Constructed Wetland;	
	Deep wells – highest in 2D (range of 14-23 mg/L) and 4D (0.8-10.5 mg/L), compared to 7D (<1.3 mg/L)	
	C wells – elevated in 5C (14-21 mg/L), 2C (9-13 mg/L) and 4C (5.3-9.8 mg/L), the rest of the C monitoring wells recorded concentrations <5 mg/L	
Chromium	Most groundwater concentrations < 0.002 mg/L, the exception is MW5C with chromium between 0.0052 - 0.012 mg/L (historical range)	
	Groundwater chromium concentration is generally lower than site surface water (such as W and E Sediment Ponds) but elevated compared to Kaikorai Stream	
Boron	Boron concentration highest in 1C (\sim 4 mg/L), 5C (\sim 3 mg/L) and 4C (\sim 2 mg/L).	
	Deep well concentrations is highest in 7D (~1.4 mg/L), 4D and 2D <0.8 mg/l	
	Boron elevated in Eastern Constructed wetland (up to 9 mg/L), estuary concentration up to 1.8 mg/L, rest <1 mg/L (note boron analysis not undertaken in GI1, GI2, GI3, GI5)	
Arsenic	Highest groundwater concentrations measured in 2D. Most results < 0.005 mg/L with the exception of 6C and 7D	
	Groundwater and site surface water concentrations in similar range	
Iron	Elevated in groundwater, in particular 1C, 4C, 2C, 6C, 5C, and deep monitoring wells.	
	Highest concentrations recorded in 2D (116 mg/L)	
	Iron concentration in groundwater an order of magnitude higher than site surface water and two orders of magnitude higher than Abbots Creek/Kaikorai Stream (GI1-GI5)	

Table 2.7 Water quality trends in groundwater

Data so.wee: Hybrid Refei-eoce Layer. Lr-12, Stats NZ, Esri, t-ERE, oam.n. Foorsquare, METI/NASA, USGS 'M:>fld I : Maxar. Created by: SChen4

Major ion chemistry

The major ion chemistry for selected water samples is shown graphically in a piper (trilinear) plot (Figure 2.14). The piper plot shows relative proportions of the major anion and cation species within the water samples and is used to display the differences (or similarities) between different water types. The plot shows the following water types:

- Sodium bicarbonate type leachate
- Sodium chloride groundwater and estuary
- Mixed water types pump station and ponds
- Magnesium bicarbonate groundwater (MW3C and MW8C)

The piper plot shows that the samples from PS3 are an intermediate water type between leachate and groundwater. This is unsurprising given that the leachate trench intercepts both groundwater and landfill leachate. The Eastern and Western Sediment Pond samples also show an intermediate chemistry. These ponds are influenced by stormwater runoff from the landfill.

Groundwater from monitoring wells potentially influenced by waste (see above review of waste distribution) show a more varied water chemistry compared to the three wells unlikely to be influenced by landfill waste (MW2C, 2D, 4D, BH103). Major ion chemistry is not available for Kaikorai Stream (SW sample sites GI1-5), however a sample from the estuary at low tide is most similar to monitoring well MW4C. The estuary sample is likely to be influenced by activities in the wider catchment.

Isotopic Analysis

Isotopic analysis of samples from pump stations (leachate/groundwater mix), selected surface water and groundwater wells is undertaken in accordance with the consent, and the results and interpretive isotopes are included in the annual monitoring report. The relative proportions of stable isotopes can change through chemical processes, such as evaporation, diffusion, or chemical reactions, and can be used to understand mixing of different water types. The following isotopes are analysed:

- Oxygen-18 in water from leachate (δ18O-H2O), relative to Vienna standard mean ocean water.
- Hydrogen-2 in water from leachate (δD- H2O), relative to Vienna standard mean ocean water.
- Carbon-13 in dissolved inorganic carbon from leachate (δ13C- DIC), relative to Vienna Pee Dee Belemite.
- Nitrogen-15 in ammonium from leachate (δ15N-NH4+), relative to atmospheric nitrogen (from October 2019).

The data below are presented relative to the local Meteoric Water Line⁵. A number of biogeochemical and physical processes can result in waters plotting above or below the MWL.

The isotopic analysis indicates the following:

- 13 C concentrations have been relatively stable with no one sampling location deviating greatly. A slight net increase in 13C concentrations can be observed in the surface water and monitoring well MW4D data sets. The enriched ¹³C data for leachate is a by-product of methane producing bacteria which use the lighter ¹²C to form CH4 (Hackley & Liu, 1996).
- The record for ¹⁵N-NH₄⁺ isotope analysis is relatively short (data available from October 2019). There is some variability in the date but no clear pattern or trends in the available data set.

The isotopic signature for hydrogen and oxygen shows

- Groundwater from MW2D and 4D plot close together and generally within ±5% of the Dunedin Meteoric Water line.
- The majority of the surface water data plots with in MWL ±5% lines.
- The majority of the leachate data points sit as a cluster above the Dunedin MWL -5% line.
- Isotopic signature of pump station samples (leachate/groundwater) suggest a mature stage of leachate methanogenesis (interaction with older waste).

 5 The Global Meteoric Water Line (MWL) describes the global annual average relationship between hydrogen and oxygen isotope (2H and 18O) ratios in natural meteoric waters (water derived from precipitation). The Dunedin MWL has been adapted for local conditions by North and Frew (2006).

The isotope data indicates that leachate is not influencing deep groundwater or surface water. The oxygen/hydrogen isotopic signature of the leachate is relatively distinct, plotting well above the Dunedin MWL compared to other samples.

Figure 2.14 Relative proportion of major ions in GILF water samples

Figure 2.15 Deviations from the Meteoric Water Line caused by various biogeochemical and physical processes (from North and Frew , 2006)

Figure 2.16 Green Island Landfill Isotope data against the Dunedin Meteoric Water Line (MWL)

PFAS

Additional water sampling was undertaken for the presence of Persistent Organic Pollutants (POP), specifically PFOS and PFOA (i.e. perfluoroalkyl and polyfluoroalkyl substances⁶). The sampling was undertaken as landfills and industrial activities are a known sources of these contaminants. Water samples were collected from the perimeter groundwater monitoring wells, surface water monitoring sites, sedimentation ponds, and the leachate collection system. The results from the groundwater samples are provided in Table 2.8 below and the location of the sampling sites are shown in **Error! Reference source not found.**, with the full set of results provided in Appendix C of the *Surface Water Report (2024)*.

Table 2.8 PFAS and PFOA Concentrations in GW monitoring wells

The concentrations of Total PFOS in the perimeter groundwater wells are at low concentrations and are below the 95% species protection limits of 0.13 ug/L, defined in the PFAS National Environmental Management Plan Version 2.0 – known as NEPM V2.0).

There was no clear pattern in the concentration and occurrence of PFAS in groundwater and the distribution of waste (particularly in areas where historic waste is present outside of the trench). The two monitoring wells (MW4C and MW7D) where PFAS has not been detected, are located in areas identified as being potentially influenced by historic waste. Conversely, monitoring wells MW1C and MW2C, which are inferred to be located outside of historic waste, contained PFAS compounds at low concentrations. The low level PFAS contamination in areas outside of historic deposition activities or in deep wells (MW4D) may relate to historic activities within the landfill and catchment prior to the installation of the leachate trench. PFAS concentration in surface water (Kaikorai Stream GI1, GI3 and GI5) are in a similar range to or greater than most groundwater samples with a significant amount of the PFAS present originating from upstream of the landfill. This is discussed further in the Surace Water report.

The concentration of Total PFOS and PFOA obtained from leachate at Pump Station 3 (PS3), which is representative of leachate mixed with groundwater, was consistent for all sampling events, with concentrations recorded at least an order of magnitude above the groundwater samples. If leachate was migrating beyond the

⁶ Synthetic chemicals found in many manufactured products

leachate trench it would be reasonable to expect that the concentrations of these chemicals in the groundwater perimeter wells would be similar to the mixed leachate PS3 sample.

Summary of water quality data

The water chemistry data shows the influence of landfill waste on groundwater quality. In areas where historical waste is known to be present outside of the leachate trench, the groundwater quality shows a mixed major ion signature with elevated contaminants.

The major ion chemistry clearly shows mixing of groundwater and landfill leachate in water pumped from the leachate trench.

The depositional environment, and relatively recent (in geological terms) change from estuarine to a freshwater setting influences the groundwater chemistry. In estuarine and shallow marine sediments elevated chloride, alkalinity and boron is likely due to the influence of sea water. Ammoniacal nitrogen (Ammoniacal-N) and iron are elevated in many of the groundwater samples, including monitoring wells unlikely to be influenced by waste. The elevated ammoniacal-N and iron in background groundwater may reflect the influence of the organic material in the estuarine sediments (KEF) and reducing conditions in the aquifer. Electrical conductivity is elevated in all samples, reflecting the influence of leachate and/or brackish water in the Kaikorai estuary.

2.5.4 Leachate levels

Leachate levels were measured in the landfill by dipping existing gas wells and groundwater monitoring wells within the landfill footprint. The leachate survey was completed in August 2022 by a site contractor. These data are included in Appendix D and shown in Figure 2.17. There is some variability in leachate measurements, this is likely due to the heterogeneity of the fill. As of August 2022, the leachate level in the centre of the landfill is on average 16 or 17 amsl over completed areas of landfill and cap, with two wells over 20 m amsl.

2.6 Other groundwater users

Bore records and bore and water take consents listed on ORC webmaps⁷ were reviewed for a 2 km radius surrounding the site. Two water takes were identified, one upgradient of the site at Blackhead quarry and the other for Maxwells Landfill (previously operated by Waste Management) (Figure 2.18).

Two bore consents were identified, one of the bore consents (RM14.355.01) is described as "proposed". The other (RM22.311.01) is described as "decommissioned". Details on the RM22.311.01 consent indicates that bore (ORC well number CE17/0153) was drilled in March 2023. The bore was drilled to a depth of 32 m (Table 2.9) and appears to have been dry.

Forty-nine bore records were identified in the area of interest, as summarised in Table 2.10. The site is not located within a mapped aquifer or groundwater protection zone⁸.

Depth	
$0 - 8$	Clay boulders, grey, dry
$8 - 12.5$	Clay boulders, brown, damp
$12.5 - 23$	Rock, grey, damp
$23 - 30$	Volcanic clay, brown, dry
$30 - 32$	Rock, grey, dry

Table 2.9 Summary log of CE17/0153⁹ located to the south of GILF

⁷ https://maps.orc.govt.nz/OtagoMaps/

⁸ https://www.orc.govt.nz/plans-policies-reports/regional-plans-and-policies/water
⁹ **View Well | Well summary (teurukahika.nz)**

Table 2.10 ORC Bore records within 2 km radius of the site

5.4 Piez2

9.6 GW22

6.3 Gas Compond GW21 104

8.7 Piez3

GW17 167

8.2 Piez4

Dunedin City Council 12547621 Project No. 12547621 **Green Island Landfill**

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Project No. **12547621** Revision No. **A 17/1/2023** Date

GW37 192 **16.3 14.1 21.8 GW15 Leach Riser GW9**

6.6 Piez1 8.8 GW11 82 GW7 88 **GW1** 105 **GW7 GW13 GW11**

16.7 GW36 142 **22.2 GW26 GW14**

Leachate head measured in landfill

nol12547621 ENV_Z001_LeachateLevel.mxd Data source: Eagle Technology, Land Information New Zealand, GEBCO, Community maps contributors. Capital source: Eagle Technology, Land Information New Zealand, GEBCO, Community maps Figure 2.17

Data Disclaimer

© 2021. Whilst every care has been taken to prepare this map, GHD (and LINZ, Environment Canterbury) make no representations or warranties about its accuracy, reliability, completeness or suitability for any particular pur

RRPP Site Boundary

² km Radius

Revision No. **A** Date **15/07/2024**

WITHIN 2 KM RADIUS

Map Projection: Transverse Mercator Horizontal Datum: NZGD 2000
Grid: NZGD 2000 New Zealand Transverse Mercato

Metres

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 \blacksquare Bore consents

otherwise) for any expenses, losses, damages and/or costs (including indirect or consequential damage) which are or may be incurred by any party as a result of the map being inaccurate, incomplete or unsuitable in any way

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Christchurch Projects Care and Christchurch 2012 (Account 2012) Account 2020 (Eagle Technology Land Information New Zealand, GEBCO, Community

3. Technical assessment

The following section documents the conceptual understanding of the groundwater system, interaction of groundwater with the landfill leachate and the leachate collection system, and interaction with surface water. This conceptual understanding, site data and investigation results, have been used to assess the effect of the proposed activities. The technical assessment includes:

- An assessment of rainfall infiltration through the current and proposed cap (HELP modelling)
- An assessment of leachate head within the fill (2D SEEP/W groundwater modelling)
- An assessment of the effectiveness of the leachate trench in intercepting leachate (SEEP/W)
- Groundwater surface water interaction

Detailed summaries of the modelling assessments are provided in and Appendix F (HELP modelling) Appendix G (Seep/W).

3.1 Conceptual model

The conceptual understanding of the groundwater system is summarised below and shown in Figure 3.1. and Figure 3.2. The model is based on published information, routine monitoring, and previous site investigations, in summary:

- Infiltration of rainfall into the landfill, generation of leachate as water comes into contact with waste;
- Migration of leachate down and outwards towards the edge of the landfill;
- Abstraction of groundwater and leachate from leachate trench, water chemistry confirms mixing of water types;
- Groundwater quality influenced by historical waste deposition; this includes areas outside of the trench (e.g MW4C). However, continuous pumping from the leachate trench maintains the hydraulic gradient and pulls the impacted groundwater towards the trench (and away from surface water);
- Stream depletion effects limited by presence of HDPE liner on stream side of trench. However, some abstraction of shallow groundwater from the outer edge of the site (and stream water) is likely to occur;
- Reduced permeability of the estuarine sediments below the landfill footprint due to compression of soft sediments.
- Upward hydraulic gradient in LKEM restricting migration of leachate into deeper layers under the trench collection system; and
- Underlying mudstone forms an aquitard limiting deeper flow paths and restricting flow to the south (southern valley) where it occurs at/near the surface.

A groundwater contour map (Figure 3.3) has been created to show the relative groundwater level across the site. Within the landfill, this shows the leachate head in the fill material. The leachate interception trench reduces the groundwater-leachate level either side of the trench as shown by the inset. The data used to create this map are included in Appendix B.

Figure 3.1 Conceptual model – North and west margins of the landfill

Figure 3.2 Conceptual model – southern valley (current and future)

Alluvium

Shallow alluvium underlain by mudstone

Dunedin City Council 12613624 Project No. 12613624 **Dunedin City Council**
Green Island Landfill

Abbotsford siltstone and mudstone no groundwater flow

> Project No. Revision No. **A**

Map Projection: Transverse Mercator Horizontal Datum: NZGD 2000 Grid: NZGD 2000 New Zealand Transverse Mercator

Groundwater contour map

22/07/2024 Date

Kaikorai stream

Data source: Eagle Technology, GNS Science emap - Eagle Technology, Land Information New Zealand, GEBCO

3.2 Review of flow paths

As discussed in Section 2.2.1 the landfill was formed on fine grained estuarine sediments and organic matter of the Kaikorai estuary. The 1942 aerial photograph shows a defined channel for Abbots Creek / Kaikorai Stream around the edge of the current site footprint. The photograph shows small channels across the estuarine sediments flowing from the centre to the southwest corner of the site. While there is no recent evidence of alternative Kaikorai Stream flow path, it is likely that the main stream channel moved around, with the general flow direction from the northeast towards the south west. Therefore, monitoring wells in the southwest corner of the site could be considered to be downgradient of the historical drainage, and any preferential flow paths (if present).

Monitoring wells in the southwest corner of the site include MW1C (4.6 m deep) and MW0C (4.9 m deep) both likely installed in the UKEM. A piezometer (BH103) installed as part of the 2022 geotechnical investigation provides further coverage along the downgradient side of the site. BH103 is screened in the LKEM, between 8.1- 10.1 m depth. The major ion chemistry (Section 2.5.3.3) of BH103 is consistent with other deep wells (2D and 4D) with very high chloride, elevated ammoniacal-N and iron, related to the estuarine organic sediments in a low DO environment. MW0C and MW1C shows some variability compared to the deep groundwater samples, with a greater proportion of sulphate, but do not show a leachate signature.

3.3 Groundwater sensitivity and monitoring

Our review of water take consents and bore records (see section 2.6) confirms that groundwater is not used for supply near the site or downgradient due to the low yielding aquifer conditions (clay and rock). On that basis, groundwater is not considered to be a sensitive receptor in accordance with the WasteMINZ landfill guidelines. In addition, the site is bordered to the south by mudstone and siltstone forming a barrier to groundwater flow.

Groundwater monitoring is undertaken along the site boundary. With monitoring wells either side of the leachate interception trench. The current and proposed groundwater monitoring is generally in accordance with the WasteMINZ landfill guidelines (2023). The absence of groundwater monitoring wells along the southern valley (section between MW9 and MW0) reflects the characterisation of the geological environment (noted as comprising extensive deposits of mudstone of the Abbotsford Formation) and the absence of a known or mapped aquifer. As discussed in Section 2.4 the hill country to the southeast of the landfill is comprised of mudstone, therefore groundwater flow towards the south and coast is unlikely. The mudstone and overlying siltstone has been the source of capping material for the landfill.

As part of the proposed leachate improvement at the landfill, DCC propose to extend the leachate interception trench into the valley to reduce leachate levels in the landfill and avoid leachate being conveyed via the surface drainage. This trench will provide additional hydraulic separation from the wider groundwater environment.

3.4 Modelling assessment

The modelling assessment utilised two different methods:

- HELP (Hydrologic Evaluation of Landfill Performance)
- SEEP/W 2D groundwater model

Rainfall infiltration through the landfill cap was assessed using the Hydrologic Evaluation of Landfill Performance (HELP) software (Berger and Schroeder, 2013). HELP 3.95D is a quasi-two-dimensional hydrologic model for conducting water balance analysis of landfills and cover systems. The model utilises weather, soil and landfill design data to account for the effects of surface storage, runoff, infiltration, evapotranspiration, soil moisture, lateral subsurface drainage, vertical drainage and leakage through soil and liners. The modelled leakage (infiltration) through the current and future landfill cap was included in the SEEP/W modelling assessment. The HELP modelling assessment is included in Appendix F. HELP can also be used to calculate leachate leakage through the base of the landfill, however this is typically applied to lined landfills. Given the complex history of the Green Island Landfill site and absence of a modern landfill liner system, it was considered that the SEEP/W was more suitable for representing the interaction between the landfill, including the leachate volumes and head level within the landfill, and the receiving environment.

The groundwater assessment included 2D modelling using Geostudio 2021 SEEP/W finite element modelling software. Modelling was undertaken to estimate the seepage into the leachate trench and to simulate the leachate head within the landfill.

Two SEEP/W cross-sections were created to model the landfill. The location of the cross-section lines is shown in Figure G.1 (Appendix G). The models were created based on a drone survey provided by DCC. Each model was initially run under steady-state conditions to simulate the interpreted baseline groundwater conditions. Models were calibrated to:

- Measured leachate head within the landfill (average level); and
- Combined pump station flows (from leachate collection trench), under dry weather conditions.

Steady state model scenarios were then run to simulate future conditions at closure, in particular, capping of the landfill and installation of a leachate collection trench in the southern valley. The Line 2 future scenarios include additional filling as outlined in the *Design Report* (GHD 2023). Details of the model set up and assumptions are provided in Appendix G. Model scenarios are included in Table 3.1 below.

Table 3.1 Model scenarios

3.4.1 Results

The results of the modelling for the base case (current) and future simulations are summarised in Table 3.2.

The modelling indicates flows to the leachate trench of approximately 1 L/s, this does not include the flows reporting to pump stations 5 to 8, which are expected to flow into the trench at a lower rate (per metre of trench) than the sections modelled. However, the results are in line with the current recorded flows (section 2.2.3).

The modelled leachate head for the base case scenarios are similar to field measurements, although the very high leachate in the centre of the landfill could not be simulated. It is likely that the high leachate levels reflect the heterogeneity of the landfill materials resulting in pockets/areas of higher leachate.

**Flow rate (per metre) multiplied by length of trench represented by each section:*

Line 1 – Perimeter trench, 400 m, Southern Valley trench (150 m)

Line 2 – Perimeter trench, 620 m, Southern Valley trench (300 m)

The relative proportion of flows from each side of the trench was calculated from the model. For Section 1, the modelling results indicate that approximately 70% of the flow to the trench comes from the landfill (and underlying groundwater) and 30% from the direction of the stream. Stream depletion rates are estimated to be < 0.1 L/s.

If the leachate trench pumps were to fail for an extended period (i.e. several weeks), the modelling shows a reversal of flow over time with flow from the landfill into the stream eventually at an estimated rate of 0.5-0.8 L/s (combined rate). This equates to approximately 0.2% of the mean flow in the Kaikorai Stream, downstream of Abbotts Creek confluence.

By undertaking a simple mass mixing model approach the potential impact on water quality in the Kaikorai Stream from a prolonged loss of the hydraulic containment can be estimated (Table 3.3). At mean flow, using the mixing model approach, the Ammoniacal-N concentration would increase by an order of magnitude when compared to existing background concentrations, whilst at low flows the concentration would be higher still.

Table 3.3 Simple mass mixing model for GW Ammoniacal-N discharge to Kaikorai Stream

It is understood that when on occasions blockages have occurred in the gravity leachate lines taking leachate to the pumpstations, , it has taken weeks for leachate levels to rise within the trench (pers comm. L Coe). The monthly water level monitoring data from the perimeter wells and leachate manholes that has been collected over the years provides a good evidentiary basis for showing that the trench has been effective in maintaining a

hydraulic barrier. It is designed to enable gravity flow of leachate along the trench to a pump station, with the ability of leachate to bypass a pump station in the event of an isolated pump shut down. In addition, it is unlikely that the entire system would fail at once for a prolonged period of time, where remedial actions were not able to be instated to maintain the hydraulic barrier. Therefore, the effects to surface water as shown by the model scenario are unlikely to be realised for short term failures.

The Section Line 2 model indicates a very small flow from groundwater to the stream (0.01 L/s) this is interpreted to be from the estuarine sediments adjacent to and under the stream bed. The stream is located ~70 m from the trench, therefore while the trench appears be effective at intercepting leachate from the landfill, it does not result in stream depletion and does not influence groundwater levels immediately adjacent to the stream.

A summary of the key finding from the modelling assessment is provided below in Table 3.4.

Feature	Summary	
Cap permeability	Decrease in leachate flow rate and leachate head level when a lower permeability cap is applied. This effect is most pronounced when a cap hydraulic conductivity (K) of 10^{-9} m/s is applied. The difference between leachate flow and head for the K 10^{-7} and 10^{-8} m/s scenarios is small.	
River side leachate trench	The trench is effective at drawing down leachate levels and intercepting flows to the stream. When the trench drainage boundary condition is "turned off" for modelling purposes, the flow rate to stream is estimated to be between 0.5-0.8 L/s. Modelled flow rates are similar to dry weather trench pump rates	
Southern valley trench	Estimated flow rates to the proposed trench are 0.2-0.3 L/s (0.03 L/s for low permeability cap scenario)	
Surface water interaction	The modelling indicates that stream water (and shallow groundwater) is intercepted by the leachate trench. Stream depletion rates are estimated to be in the order of 0.1 L/s for Section 1. Stream depletion rates are higher in the sea level rise scenarios due to the higher water levels.	
	However, the Section 2 model indicates flow (at a very low rate, 0.01 L/s) into the stream from shallow groundwater adjacent to the stream. This is much lower than the modelled flow to the stream $(0.3\n-0.5 \text{ L/s})$ if the trench was "turned off" showing that the trench is effective at intercepting leachate. The small inflow (0.01 L/s) to the stream under operating conditions is interpreted to represent groundwater seeping from the stream bed and adjacent river bank. Due to the distance of the trench to the stream, the operation of the leachate trench does not influence groundwater levels immediately adjacent to the stream.	

Table 3.4 Modelling summary

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4. Assessment of effects on the environment

4.1 Introduction

The proposed filling and closure of the Green Island Landfill has the potential to affect water quantity and quality. An assessment of effects addressed in this technical assessment includes:

- Effects to groundwater levels and flow;
- Effects to surface water flows;
- Effect of climate change and sea level rise; and
- Effects to water quality.

The water quality assessment is focussed on groundwater quality with consideration of surface water interaction. Surface water quality is primarily addressed in *Surface Water Technical Assessment* (GHD, 2024).

4.2 Effect to groundwater and surface water levels and flows

The leachate collection system operates by drawing down water levels in the trench, this intercepts any leachate flowing from the landfill but also draws groundwater from the area outside of the trench. It is likely that this groundwater is hydraulically connected with surface water in the Kaikorai Stream, with the potential for groundwater abstraction to have a stream depleting effect.

The modelling assessment presented in Section 3 and Appendix F, indicates that approximately 30% of the water pumped from the leachate trench is derived from groundwater / connected surface water on the outside of the trench, in areas where the trench is close to the stream. This volume is estimated to be <0.5 L/s for the entire trench length. For areas further from the active stream channel, eg PS3, the modelling indicates no effect of the leachate trench operation on stream flows.

Overall, the assessment indicates the effect of the leachate trench abstraction on surface water flows is negligible when compared to the stream flows, even during low flow conditions, and volume of water in the estuary.

The underlying KEF and Abbotsford Formation are not used for groundwater supply. Therefore, the abstraction of groundwater and localised reduction in groundwater levels around the landfill perimeter does not affect any groundwater users.

4.3 Effect of climate change and sea level rise

Modelling was undertaken to assess the effect of higher river levels on the leachate collection system. The modelling incorporated a higher road level (as recommend in the Design Report) and a higher river level (increase by 0.5 m). The model results indicated slightly higher inflows to the leachate collection system, this increase is well within the operating range of the leachate collection system, which can accommodate much higher stormwater flows.

4.4 Effects to water quality

This assessment has shown the perimeter leachate collection system is effective at creating a hydraulic barrier and intercepting leachate flowing from the landfill. It also draws groundwater from outside of the gravel leachate trench, this is of particular importance in areas where waste is present outside of the trench (from historic activities). This abstraction prevents the movement of potentially contaminated groundwater from outside the leachate trench, into surface water. As discussed in the *Surface Water Assessment,* surface water monitoring shows that there are no discernible adverse effects on surface water quality from the existing landfill activities.

The review of ORC water takes and bore records did not identify any groundwater users close to or downgradient of the site. Furthermore, the ORC has not mapped the area as a groundwater resource. Therefore, on that basis groundwater is not considered to be a sensitive receptor.

The modelling assessment was used to assess the effect on groundwater flow rates and flow paths should pumping from the leachate trench cease for a prolonged period resulting in discharge to surface water. This modelling scenario is considered very unlikely in reality, as there would be a time lag for leachate levels to rise in the trench before the modelled flow rate was achieved. Furthermore, it is considered unlikely that all pump stations would be out of action for an extended period given historical performance and additional mitigation measures that are proposed in the *Design Report*. However, for the purposes of this assessment, if leachate migration to surface water was occurring (at a rate of $0.5 - 0.8$ L/s) we would expect to see a measurable change in water quality between the surface water monitoring points GI2 and GI3 (during low flow conditions) due to the very high contaminant load in the landfill leachate (as described in section 3.2). Instead, the water quality at the downstream GI3 is generally better than GI2. This further supports the conclusions of modelling assessment that the leachate trench is effective at intercepting landfill leachate.

The proposal includes the installation of a leachate collection trench in the southern valley, this will intercept leachate flowing from the southern side of the landfill. However, there will still be a gap in the leachate trench between MH8 and PS9. This area of the landfill sits directly on a ridge of Abbotsford Formation. The mudstone forms an effectively impermeable barrier to flow, therefore leachate migration off site is unlikely.

4.4.1 Effects on recreational use

While the post closure use of the site is still to be confirmed, the applicant does not intend to re-use or apply impacted groundwater to land (via irrigation). Therefore, the groundwater quality will not impact any future recreation use of the site and surrounding area. With regards to surface water quality, The Human Health Risk Assessment (HHRA) (GHD,2024b) concluded that discharges from the site into the receiving environment of the Kaikorai Stream generally represent a low risk to human users of the waterway and the aquatic environment.

4.5 Effects on other groundwater users

The surrounding area is not used for groundwater supply due to the low permeability geology, with the only two consented groundwater takes (within a 2 km radius of the site) located upgradient of the site (quarry and landfill). Groundwater is not utilised to the south of the site due to low permeability ground conditions. As noted above, the effect of the proposed dewatering activities of groundwater are expected to be limited to the area within the leachate perimeter trench with no effect on any other groundwater users.

5. Recommendations

5.1 Monitoring

It is recommended that a condition of consent for the provision of a Groundwater Monitoring Plan is included, which would include the following:

- Details of all monitoring locations (groundwater and surface water),
- Details of the monitoring well construction details, elevation, screened interval
- A detailed sampling and analysis plan
- The practice and procedures for groundwater monitoring will be included in the Landfill Development Management Plan (LDMP)

For both groundwater monitoring and surface water monitoring it is recommended to continue with the monitoring programme in line with the current consent conditions, with the following exceptions:

- It is recommended that isotopic analysis is removed as a consent condition. The current consent requires quarterly sampling for analysis of environmental isotopes. It is considered that isotopic analysis does not significantly improve the understanding of the groundwater/surface water system, and given the long delay (months) between collection of samples and reporting of results, chemical analysis is likely to provide a more timely indication of leachate mobility.
- It is recommended that the requirement to install a deep groundwater well within the landfill is removed as a consent condition. The installation of a deep well has the potential to create preferential flow paths from fill to the underlying geology.
- It is recommended that deep well BH103 is added to the routine monitoring schedule to provide geographical coverage of the (historic) downgradient side of the site.

The following total peak and average daily flow rates are proposed as a consent condition for the leachate collection system. These maximum rates allow for both groundwater and diverted stormwater during potentially extended periods of rainfall:

- Peak rate 20 L/s $(1,728 \text{ m}^3/\text{day})$
- Average rate 5 L/s (432 m³/day or 157,680 m³/year).

It is recommended that the following water level and quality monitoring is undertaken (Table 5.1).

A comparison to the existing monitoring schedule is outlined in Table 5.2. The proposed changes simplify the current monitoring program, with all water types having the same quarterly and annual analytical suites. Some analytes have been removed from the schedule, as these analytes (i.e. faecal coliforms, volatile fatty acids) were considered not to be useful for detecting leachate migration into the environment.

Frequency	Measurement/Analyte	Locations
Monthly	Groundwater levels	$A/B/C/D$ wells, BH103, pump stations and manholes.
Quarterly ¹	рH Electrical Conductivity	- C and D wells - BH103
	Dissolved oxygen Boron	- Representative sample from the leachate trench (PS3)
	Nitrate Nitrogen Ammoniacal Nitrogen Chloride	- Surface water (GI1, GI2, GI3, GI 5 and estuary) within three hours of low tide
	PFAS/PFOA3	- Western sediment pond - South western pond - Eastern sediment pond

Table 5.1 Recommended monitoring

¹Reduced frequency to 6 monthly, two years post closure

2Metal analysis – dissolved metals in groundwater and leachate trench (PS3), total metals in surface water samples.

³Reduced frequency to annual after 3 years

Water quality results are to be documented and reviewed and assessed on receipt, with full analysis and reporting in the annual monitoring report due 1 October every year. The annual monitoring report is to include a discussion on any water quality trends and the effectiveness of the leachate trench in intercepting landfill leachate.

Post closure it is recommended that the monitoring programme is reviewed and updated to reflect the changes to the landfill.

Table 5.2 Comparison between current and proposed monitoring schedule

6. Conclusions

This technical groundwater assessment has involved a review of historical data, including the historic distribution of waste, site monitoring data, and recent site investigation data. This information was used to inform a groundwater modelling assessment to estimate the seepage into the leachate collection trench and to simulate the leachate head within the landfill. In summary:

- The leachate collection trench is effective in intercepting landfill leachate and drawing groundwater from areas of waste outside the trench;
- The chemical signature of waters pumped from the leachate trench show that it is a mix of groundwater and leachate;
- The future modelling scenarios indicate a small reduction in leachate head in recently capped area (northern part of the landfill) and an increase of leachate head in areas receiving additional fill;
- Modelling indicates that the leachate trench in the southern valley will intercept leachate coming from the landfill and aid to lower leachate levels within the landfill and enhance leachate collection;.
- Potential for a small increase in dry weather flows in the leachate collection system from the additional fill placement and installation of the southern valley trench;
- With the continuing operation of the leachate collection system (including the extension of the trench into the southern valley), and maintenance of the groundwater hydraulic barrier, no discernible effect on surface water quality is expected; and
- The effect on surface water flows from groundwater abstraction is negligible.
- Groundwater is not considered to be a sensitive receptor as groundwater is not utilised for supply in the surrounding area due to the low permeability geology.
- There is no evidence of groundwater flow occurring to the southeast, under the landfill and towards the coast.

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8. Limitations

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GHD has relied on information from a number of sources, including but not limited to the following:

Environmental Impact Assessment (Beca, 1992)

Green Island Landfill Leachate Collection Trench Geological Report (Barry J Douglas Geological Consultants, 2002 – herein referred to as BDGC, 2002)

Green Island Landfill Gas Management Project. Stage 1: Investigation Works. URS, 2007.

Bund stability report (T&T, 2020)

Leachate collection summary (DCC) and associated plans

Claymine Assessment (DCC) and associated plans

GILF & RRRP master plan (Stantec 2020)

Green Island Annual monitoring reports prepared by GHD, 2022

Appendices

A-1 Hydraulic conductivity testing

Rising and falling head tests of existing monitoring wells and new bore holes were undertaken in November and December 2022 to estimate the hydraulic conductivity of the underlying shallow aquifer. Up to two sets of rising and falling head tests were completed in each well, however due to the slow groundwater response in the low permeability sediments, less tests were completed in some wells. The raw data was initially processed and reviewed graphically and then, if test results were considered suitable (e.g. not too many fluctuations, and suitable measured response and recovery time), data were interpreted with Aqtesolv software using Bouwer and Rice (1976) and Hvorslev (1951) solutions.

The hydraulic conductivity estimate for monitoring wells MW2D is based on historical data from 1997, when the well was purged and groundwater levels recovered over a few weeks.

The results of the hydraulic conductivity testing are summarised in Table A.1, full results are provided in Table A.2 with analysis plots following.

Geology	Monitoring wells*	Hydraulic Conductivity Range (m/s)	Adopted Hydraulic Conductivity (m/s)
Upper Kaikorai Estuary Member (UKEM)	MW1C, MW5C, MW6C	8.4 x 10 ⁻⁷ to 2.8 x 10 ⁻⁶	1×10^{-6}
Lower Kaikorai Estuary Member (LKEM)	BH100, BH101, BH103, BH108, MW2D, MW8C	6.2 x 10 ⁻¹⁰ to 3.3 x 10 ⁻⁶	1×10^{-7}
Abbotsford Formation (AM)	BH104	$<$ 1 x 10 ⁻⁹	1×10^{-9}

Table A.1 Summary of hydraulic conductivity testing

*Logs not available for older MW monitoring wells, inferred geology based on recent drilling investigation.

Wells BH102, MW0C, MW2C, MW3C, MW4C, MW4D likely screened over two layers

Test in BH104 does not recover fully (<85% recovery) – results approximately only.

Table A.2 Hydraulic conductivity testing

**FHT/RHT = Falling/rising head test undertaken using displacement slug*

B-1 Bore locations and water levels

Notes:

1 New Zealand Transverse Mercator

2 Otago metric datum

³New Zealand Vertical Datum. Locations on monitoring wells lines 1-9 converted from OMD based on correction factor of -(0.388 m + 100 m), from height of mark AG1B **AG1B: W 6 (linz.govt.nz)**

4Locations and elevation not surveyed, measured with hand held GPS (approximate only)

@ 2011. Whilst every care has been taken to prepare this figure, GHD make no representations or warranties about its accuracy, reliability, completeness or suitability for any particular purpose and cannot accept liability

Note: Flood event during the July monitoring round - dilution is likely cause of reduced concentrations in surface water sampling locations

Figure C1-1 Ammonical Nitrogen

Note: All arsenic concentrations at locations GI1, GI2, GI3 & GI5 were less than the detection limit

Note: Flood event during the July monitoring round - dilution is likely cause of reduced concentrations in surface water sampling locations

Figure C1-2 Arsenic

Note: Boron analysis not undertaken in samples GI1, GI2, GI3, GI5

Note: Flood event during the July monitoring round - dilution is likely cause of reduced concentrations in surface water sampling locations

Figure C1-3 Boron

Note: Flood event during the July monitoring round - dilution is likely cause of reduced concentrations in surface water sampling locations

Figure C1-4 Chromium

Figure C1-5 Iron

Analytica Laboratories Limited Ruakura Research Centre 10 Bisley Road Hamilton sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd

Level 1, Bing Harris Building, 286 Princess Street, Dunedin Submitted by: Dunedin 9016

Attention: Dusk Main Phone: 03 479 9481 Email: hayden.erasmus@ghd.com

Sampling Site: Green Island Landfill

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories (or at the subcontracted laboratories, when applicable). Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

Lab Reference: 24-19939

Date Received: 22/06/2024
Testing Initiated: 24/06/2024

Date Completed: 15/07/2024 Order Number: 12613624 Reference: 12613624

Testing Initiated:

Courtney Deavoll

Biochemical Oxygen Demand

Anion/Cation Suite

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited.

Anion/Cation Suite

Elements in Water (Soluble)

Carbon in Water

Method Summary

Sharelle Frank, B.Sc. (Tech) Technologist

Sandra Mathews, B.Eng. Technologist

Thara Samarasinghe, B.Sc. Technician

Bhumika Patel
Sample Reception Technician

Green Island Landfill

Landfilling History and Targeted Contaminated Land Assessment Report

Dunedin City Council

7 March 2023

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Document status

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Attachment A Historical Plans and Figures Attachment B Borehole Logs Attachment C Analytical Results Tables Attachment D Laboratory Reports

1. Introduction

1.1 Background

Existing Otago Regional Council (ORC) resource consents for the operation and management of the Green Island Landfill (the Site), expire in October 2023. It is understood that Dunedin City Council (DCC) is considering options for disposal of waste beyond this date, and eventual closure scenarios for the Landfill.

The Green Island Landfill (the Site) is a municipal landfill facility situated on Taylor Street, to the west of Brighton Road, approximately 10 km southwest of the suburb of Green Island and central Dunedin. A Site Location Plan is provided as Figure 1, overleaf.

GHD Limited (GHD) has been engaged, as part of a wider project to prepare planning documents for application to consent future landfilling and closure, to undertake a review of the history of landfilling at the Site and an assessment of the contamination status of soil recovered during drilling works being undertaken as part of a concurrent GHD geotechnical investigation. The geotechnical investigation was focused on the perimeter of the landfill, with boreholes located mainly around the leachate collection drain / access track area. Further details of borehole locations are provided in Section 4 of this report.

The findings of this report will provide contaminated land status context for both the geotechnical and hydrogeological assessment repots and help inform the design report and consenting process.

1.2 Purpose and Scope of the Environmental Assessment

The purpose of this piece of work is to improve the understanding of the history and extent of landfilling at the Site and assess soil contaminant status in the geotechnical investigation boreholes. This work helps inform the technical reports being prepared as part of the Site's continued operation and closure works and consent application strategy.

The following scope of works was undertaken:

- A review of historical aerial photographs for the Site sourced from Retrolens and DCC.
- A review of historical site investigation reports, Site plans, and compliance reports for the Site.
- Collection of samples from soil cores retrieved during the GHD geotechnical investigation.
- Submission of selected samples to the laboratory for analysis of identified contaminants of concern.
- Comparison of analytical results to adopted guidelines and standards.
- Preparation of a report outlining the history and extent of landfilling at the Site and discussing the findings of the site investigation.

1.3 Assumptions

GHD has made the following assumptions during the preparation of this report:

- Information obtained from third parties and DCC is complete and accurate.
- That the Site will remain in commercial / industrial land use until closure and thereafter will be used for recreational purposes.

Figure 1: Site Location Plan – Green Island Landfill

1.4 Limitations

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The opinions, conclusions and any recommendations in this report are based on assumptions made by GHD described in this report (refer section(s) 1 through 7 of this report). GHD disclaims liability arising from any of the assumptions being incorrect.

The opinions, conclusions and any recommendations in this report are based on information obtained from, and testing undertaken at or in connection with, specific sample points. Site conditions at other parts of the site may be different from the site conditions found at the specific sample points.

Investigations undertaken in respect of this report are constrained by the particular site conditions, such as the location of buildings, services and vegetation. As a result, not all relevant site features and conditions may have been identified in this report.

Site conditions (including the presence of hazardous substances and/or site contamination) may change after the date of this Report. GHD does not accept responsibility arising from, or in connection with, any change to the site conditions. GHD is also not responsible for updating this report if the site conditions change.

GHD has prepared this report on the basis of information provided by Dunedin City Council and others who provided information to GHD (including Government authorities), which GHD has not independently verified or checked beyond the agreed scope of work. GHD does not accept liability in connection with such unverified information, including errors and omissions in the report which were caused by errors or omissions in that information.

2. Site Description

2.1 Site setting

The Green Island Landfill (the Site) is a municipal landfill facility situated on Taylor Street, to the west of Brighton Road, approximately 10 km southwest of the suburb of Green Island and central Dunedin. The facility is currently managed and operated by Waste Management Ltd., on behalf of DCC who own the landfill.

The Site is located in a reclaimed wetland area within the Kaikorai Estuary, which is part of the larger Kaikorai Catchment; a 55 km² area bounded by the Kaikorai Stream and Abbots Creek and the topography of Chain Hills, to the northwest and north, Kaikorai and Round Hills to the north and northeast and Saddle Hill, to the west.

The Kaikorai Stream historically ran through the Site but was later diverted along the western boundary of the Site to run in a southwest and southerly direction, towards the Kaikorai Lagoon and ultimately the sea. The stream forms the northern and western limits of the landfill before flowing into the Pacific Ocean near Waldronville.

The landfill is approximately 38 hectares in size and is delineated by the legal descriptions in Table 1 below.

Table 1: Legal Descriptions for Green Island Landfill

Several activities are currently being undertaken within the boundaries of the landfill including municipal waste disposal, compost production, liquid waste and sludge disposal alongside the operation of a waste transfer station and a recycling centre. A Site layout plan is presented as Figure 2, overleaf.

As can be seen on Figure 2, the working face is currently located towards the centre-south of the landfill, the composting area is located near the eastern boundary and the waste transfer station and recycling station in the north-eastern portion of the landfill. Landfilling is complete over the northern and eastern portion and continues over the remainder of the Landfill.

A network of landfill gas collection wells is installed at the landfill. Additional wells have been installed over time with the development of the landfill. The configuration of the wells changes over time to allow for improved gas collection and collection of gas from areas of new waste. The gas is piped to the Green Island Waste Water Treatment Plant (GIWWTP) for use in the generation of electricity.

A leachate collection trench runs around the perimeter of the northern, western and southern boundaries of the Site. An access track is generally aligned over the top of this trench along much of its length.

Two stormwater sedimentation ponds are located on the landfill, one on the southwestern site boundary (West Pond) and one on the northeast site boundary (East Pond), see Figure 2 overleaf.

2.2 Contaminated land (HAIL) status

The requirements of the Resource Management (National Environmental Standard for Assessing and Managing Contaminants in soil to protect Human Health) Regulations (NESCS, 2011) applies when a selected activity on a piece of land where an activity or industry on the Ministry for the Environment (MfE) hazardous activities and industries list (HAIL) is, has, or is more likely than not to have occurred. Activities such as soil disturbance, subdivision and change in land use are regulated by the NESCS

The Site is known to be part of a landfill and as such is categorised as having HAIL activity G3 (landfill sites) occurring on it.

Figure 2: Site layout Plan

3. Landfill History

3.1 Site history

A landfill has been present on the eastern side of the Kaikorai Estuary since 1954. It is understood that waste filling began on the eastern side of the upper Kaikorai Estuary and by the late 1970s, the Site had become the main landfill for Dunedin.

Unregulated and uncontrolled landfilling occurred at the Landfill until the 1980s, when DCC began to manage waste disposal activities through a national planning approach for the area. The landfill was granted resource consents under the RMA in 1994 and a combined leachate interception trench and collection system was constructed at the Site. This comprised nine (9) pump stations interconnected via a gravel-filled trench with an inbuilt perforated collector drain located around the landfill toe and was retrofitted around the majority of the perimeter of the landfill.

This pump network is set up to maintain a hydraulic gradient towards the trench, minimising the amount of leachate migrating beyond the interceptor trench. The trench is not currently present along the southern boundary of the landfill, between MW0 and PS9, as shown on figure 006116-19-01 Green Island Landfill – Leachate collection and environmental monitoring system, 2004, included in Attachment A. A stormwater interception ditch is located along this boundary which is channelled towards pump station PS1. This gap in the system ws to allow planned placement of waste over this area. In the last few years DCC have decided not to place waste in this area The interception trench allows for the leachate to be collected and discharged to the Green Island WWTP, via a pipeline, located to the southwest of the landfill.

A network of groundwater / leachate monitoring wells was installed in a series of lines crossing perpendicular to the interception trench, to monitor groundwater / leachate levels across the trench to confirm hydraulic containment of the shallow groundwater. This network consists of both shallow and deep monitoring wells and each line is located approximately halfway between each pump station.

A schematic cross section plan of the landfill and the location of the leachate collection drain and monitoring well arrangement is presented in Figure 3 below and in figure 006116-19-01 included in Attachment A.

Figure 3: Schematic cross section through an example monitoring well line and the leachate collection trench (Source DCC Landfill Annual Survey Plans – July 2004 Sheet No. G11).

An HDPE liner was historically placed between the leachate interception trench and the Kaikorai Stream to minimise flow from the stream and groundwater migrating eastwards towards the landfill. During this time (HDPE liner placement), a clay bund was also installed around the site boundary to contain both the landfill (deposited waste) and leachate.

3.2 DCC Reports and Plans

A selection of reports provided to GHD by DCC have been reviewed to understand the history of the landfill and its evolution. These are discussed in the following sections.

3.2.1 Beca Steven EIA

In 1992 Beca Steven prepared an Environmental Impact Assessment¹ (EIA) report for the landfill. This report states the following:

Section 4.1.2 The Site

In 1954, an industrial landfill was started on the eastern side of the upper Kaikorai Estuary. At the time, the Kaikorai Stream was still seriously contaminated from industrial dumping, storm water runoff and household waste. There was little thought given to the impact that such an activity might have on the wetland environment in the vicinity, although much of the wetland area was farmed. The same reasons that made the area ideal for industrial development (i.e. close to the city but out of the Otago Harbour catchment), also made it ideal as a refuse site. In the early years of the landfill, as was common practice around the country, little attention was paid to what was dumped or buried on the site.

By the late 1970s Green Island had become the main landfill for all of Dunedin, taking in an average of 58,000 tonnes of refuse annually.

Currently the landfill at Green Island and the adjacent private Maxwell's landfill just across the Kaikorai Stream (started in 1968) cover over 30% of the total estuarine landscape of the valley to a maximum depth of 10 metres. Adjacent to the Green Island Landfill the Kaikorai Stream has been channelled behind bunding for the purposes of the drainage and protection of farmland. One consequence of this has been that the bunded areas form in effect oxidation ponds helping to control and treat leachates.

Section 4.2.2 Existing Features of the Site

The Green Island landfill is situated at the head of the Kaikorai Estuary and is separated by a bund constructed of in situ materials from the Kaikorai Stream. The photograph in Figure 4.2 (see below) shows the position of the site within the locality (Maxwell's landfill is to the left and Green Island DCC landfill to the right).

¹ Beca Steven (1992) Environmental Impact Assessment of the Extended Green Island Sanitary Landfill. October 1992.

The site itself is essentially a flat area on top of a raised platform. The flat area presents an untidy appearance due to the presence of windblown rubbish over much of the site and the limited "finishing" of the filled areas which has taken place.

The primary uses of the site are for refuse processing and disposal by land filling.

A prominent use of the site is the stockpiling of car bodies prior to periodic shipment to Auckland for scrap metal recovery.

Section 4.3.4 Refuse Acceptance Policy

Normal domestic and commercial wastes are accepted automatically and whether they contain hazardous or special wastes is left to the disposer to identify, although the landfill operator can inspect any load if there is reason to suspect the material may contain special or hazardous wastes.

Routine special wastes such as dead animals and small volumes of offal are accepted and disposed of by burying at the toe of the landfill.

Section 4.3.6 Surface Water Management

Streams and watercourses approaching the site are diverted around it - one to the south of the landfill which runs outside the southern bund, and one which is channelled down the eastern side of the landfill and discharged to Kaikorai Stream. The watercourse to the east is not well separated from the landfill and contamination of it by leachate occurs over the lower reaches.

Surface water from the area of the completed existing landfill is collected in channels and discharged away from the face. However many of the channels are not well formed or graded, and are affected by landfill settlement. While collecting and preventing some surface water from entering the landfill they are not totally effective. Nevertheless the upper levels of the landfill are relatively dry.

None of the upstream catchments to the south and east are sufficient in size to generate large flows and the opportunity for them to flood the site is virtually non-existent.

To upgrade stormwater control for the existing landfill operation at Green Island, the following mitigating works were recommended in the Existing Green Island Landfill EIA:

- Constructing and operating two sedimentation ponds within the landfill site
- providing a system of surface drains to safely convey stormwater within the site.
- providing for the safe conveyance of stormwater around the site without causing erosion of streambanks or the landfill. This is incorporated in the proposed construction of the outer perimeter face.
- implementing sound sediment control practices.
- conducting a monitoring programme on the discharge from the sedimentation ponds.

Section 4.3.7 Leachate and Groundwater controls

As described in the Existing Green Island Landfill EIA the recommended method of collecting leachate contaminated groundwater is in a gravel filled trench around the landfill perimeter with a perforated collector drain. Pump stations at about 200m spacing will draw down the groundwater table so that an inward hydraulic gradient is formed. Leachate contaminated groundwater from under the landfills will be intercepted by the trenches which will also draw water from the stream side and the discharge will be piped to the Green Island Wastewater Treatment Plant.

However, because the contamination of groundwater has most likely been confined to the immediate vicinity of the landfill little has travelled beyond the bunded area as evidenced by the water testing carried out to date. No groundwater wells are used in the area. The improved groundwater quality after collection of the leachate, will not be a noticeable benefit in itself, but indirectly will improve the Kaikorai Stream by preventing leachate discharge to it.

Summary

The following is understood:

- Development of the landfill was ad hoc for two to three decades before a proper management plan was put in place.
- Municipal, commercial / industrial, and special and hazardous waste were accepted at the landfill.
- Prior to the installation of the leachate collection trench in 1994, there was very little historical control of leachate collection or discharge and as such leachate was discharging into the Kaikorai Stream.
- Prior to the construction of the sedimentation ponds, stormwater management was historically very limited.
- The stream is contaminated, as is much of the catchment, due to historical commercial and industrial activities upstream of the Site.

3.2.2 Trench installation report

According to the Green Island Landfill Leachate Trench Geological Report², the landfill was developed in the upper reaches of Kaikorai Estuary without modern liner and drainage systems beneath the waste and fill to control leachate migration to the underlying sediments and groundwater.

In 1994, a 1,757 m long, gravel filled trench was constructed around most of the perimeter of the landfill to intercept groundwater and leachate flowing from the site. Pumps were installed in the trench to collect leachate / groundwater from the trench and to maintain a hydraulic barrier to manage discharge of leachate to the surrounding environment.

This Report included photographs and stratigraphic logs describing the soil conditions encountered during the trench excavation. An assessment of stratigraphy in this report observed that landfill refuse and/or fill material was the upper most material at 41 of the 76 locations logged and that the maximum thickness of landfill material observed was 2.6 m. The figure with locations of the soil logs was not included in the document, however the locations of the different trench profile logs can be interpreted from their position in relation to individual pump station, manhole and air vent locations.

 2 Green Island Landfill Leachate Trench Geological Report, Barry J. Douglas Geological Consultants, August 2002

A summary of the locations of where landfill material was encountered, its thickness and composition is provided in Table 2. Please refer to Figure 006116-19-01 in Attachment A for trench profile locations (logs).

Trench profile number	Location	Thickness of landfill (m)	Comments
17	6m north of PS3	0.5	Earthfill intermixed with bricks and shells, black leachate stained refuse
18	33m north of PS3	0.1	Earthfill intermixed with refuse
20	MH ₃	1.5	Earthfill and mixed refuse
21	29m north of MH3	1.6	Mixed refuse
22	36-42m north of MH3	2	Mixed refuse
23	17-18m south of PS4	0.8	Mixed refuse
25	30.5m north east of PS4	1.4	Mixed refuse
26	38m north east of PS4	1.25	Mixed refuse
27	42m west of AV4	2.1	Mixed refuse
29	0.5 west of AV4	2.6	Mixed refuse
30	AV4	2.6	Mixed refuse
31	4m east of AV4	0.65 (north wall) 2.6 (south wall)	Landfill
32	13m east of AV4	0.5	Brown earthfill intermixed with refuse
34	32m east of AV4		Landfill removed during and after trench construction
42	36m east of PS5	0.4	Intermixed refuse and earthfill
43	49m east of PS5	0.3	Brown earthfill intermixed with refuse
51	MH ₆	2.0	Landfill
52 (A)	Beneath access road (46m north west of PS7)	-6.0	Mostly earthfill with minor intermixed solid rubble (this material was placed in the 1970's-1990's period)
52 (B)	40m north west of PS7	3.75	Mostly earthfill with minor intermixed solid rubble

Table 2: Landfill material location, thicknesses and composition

Notes:

PS – Pump station

MH – Manhole

AV – Air vent

As can been seen for the above table, refuse material is present on the outside of the leachate collection trench from north of pump station PS3 on the western side of the landfill, to Manhole MH8 on the eastern side of the landfill. The greatest thickness of material was present between Pump station PS4 through to Air vent AV4.

3.2.3 Landfilling contour plan

GHD reviewed a figure depicting the estimated locations of the tip faces at the landfill from 1964 to 2001. This plan has been provided to GHD by DCC and its provenance is unknown, though is likely hand drawn based on aerial photographs review and site history knowledge of previous site managers.

In general, this figure supports other documents and photographs reviewed, in that it shows landfilling operations starting in the southeast portion of the site and advancing to the north and west over time. The historical tip faces appear to be estimates, and the figure does not show landfilling activities outside of the leachate trench. A copy of this figure is included in Attachment A. The historical placement of waste pre installation of the leachate interception trench is discussed further in Section 3.4.

3.2.4 Masterplan 2021

A landfill gas (LFG) masterplan for the landfill was prepared by Tonkin and Taylor in 2021³ .

The report states that in July 2020, there was approximately 4.8M tonnes of waste in the landfill. The history of landfilling at the landfill is summarised in Table 3. The purpose of including this history was to understand the potential generation of landfill gas from historically placed waste. Although the landfill was operational since 1954, tonnage data was first recorded in 1964.

Table 3: Landfilling history (source – Tonkin and Taylor Landfill gas masterplan report (2021))

The report states that the landfill has received a mix of municipal waste, commercial and industrial waste, and construction and demolition waste.

This report does not provide a plan showing the extent of the landfilling.

³ Tonkin and Taylor (2021)Landfill Gas Masterplan, Green Island Landfill. Job number 1008787.5010.v2. Dated May 2021.

3.3 Historical Aerial Photographs

As part of this report a review of historical aerial photographs from the online repository Retrolens was conducted to assess the landfilling sequencing and historical extent of the landfill.

In general, this review shows that landfilling began at the site before 1958. Landfilling then progressed from the southeast portion of the current property to the north until the eastern portion of the property was covered with fill by 1967. From there, filling advanced northwards and to the west over time. The approximate current landfill footprint was filled by 2000. The 2013 aerial photograph generally shows the Site in its current configuration.

In addition to the aerial photographs obtained from Retrolens, DCC provided GHD with aerial photographs of the Site from 1994 and 1999. These two aerial photographs are of a higher resolution than those obtained from Retrolens. A summary description of the aerial photographs reviewed is provided in the following sections.

3.3.1 1942 Aerial Photograph

No evidence of landfilling is apparent in the 1942 aerial photograph. There is land disturbance evident on the southern boundary, possibly winning material from the hillside for creation of vehicle access to the stream-edge to construct boundary drains.

There appears to be a drain / trench that runs on the inside of the Kaikorai Stream and continues around the boundary of the existing landfill property. There also appears to be another drain or fence which crosses the Site diagonally from northwest to southeast across the wetland.

A raised spur can be noted on the southern site boundary.

(Source – Retrolens)

3.3.2 1947 Aerial Photograph

No evidence of landfilling is apparent in the 1947 aerial photograph.

It can be noted that there appears to be a spur of raised ground in the southern portion of the Site with trees at the top. It appears that this portion of the Site increases in height towards the west.

(Source – Retrolens)
3.3.3 1958 Aerial Photograph

There appears to be an access road into the future landfill at the southeast corner of the current facility. At the end of the access road, it appears that landfilling or initial site development has commenced at the site. It is likely that the area on the southern boundary is being raised / built up to create a road to allow access into the landfill.

⁽Source – Retrolens)

3.3.4 1962 Aerial Photograph

Landfilling has extended westwards along the southern boundary. A structure, maybe a containment bund, is present on the northern boundary of the landfilling area.

⁽Source – Retrolens)

3.3.5 1970 Aerial Photograph

Landfilling has continued on the southern boundary of the site and has extended to the north to the edge of the boundary drain. The area of the higher spur of land from the 1940's and 1950's is clearly obvious as a different colour (lighter) along the southern boundary, indicating that this area continues to be used as a source of soil material used during landfilling works.

The southeast corner of the landfill (where filling began) has vegetation growing on it, indicating that filling in that area was complete by 1970.

(Source – Retrolens)

3.3.6 1978 Aerial Photograph

Landfilling continues to the north, west and to the northwest. Along a portion of the northern boundary of the landfill , fill has been deposited as far as the inner drain and appearing to infill it. Again, the area of the higher spur of land , visible in the 1940's photographs, appears to have soil disturbance activities being undertaken on it and likely continues to be used as a source area for soil to be used during landfilling activities (cover material etc.).

A portion (southern) of the drain that ran northwest to south east across the Site has been infilled.

(Source – Retrolens)

3.3.7 1985 Aerial Photograph

Filling appears to have progressed north-eastwards across the footprint of the landfill area, with minimal expansion to the west. Waste appears to be have infilled the boundary drain along the central portion of the northern boundary and likely extend beyond it towards the Kaikorai Stream in isolated areas.

Maxwell's Landfill is present on the opposite side of the Kaikorai Stream.

(Source – Retrolens)

3.3.8 1990 Aerial Photograph

The access road was moved to the northern portion of the Site since 1985. Areas of revegetation are visible on the southeast portion of the site. Fill has expanded to the west especially in the northern half of the landfill. It appears that the boundary drain has been infilled further westwards along the northern boundary with evidence of soil disturbance / landfilling on the northern side of it in one area.

(Source – Retrolens)

3.3.9 1994 Aerial Photograph

In the 1994 aerial photograph, an access road has been constructed around the northern and western boundary of the Site (along Kaikorai Stream). This access road generally follows the current leachate interceptor trench. Access to the Site is from the east into the northeast corner of the Site, which is the current configuration.

Active filling at the landfill appears to have been concentrated on the western portion of the site and particularly onto the wetlands at the southwest corner. Several support buildings appear to have been built since 1990. Revegetation is ongoing in the southeast quadrant of the site.

It appears that that access track / leachate collection trench has been constructed along the western and northern boundaries. The trench / drain which had run along the southern boundary has been infilled.

It appears that the Eastern and Western sedimentation ponds are under construction and that the north-eastern and south-eastern ponds have been construction.

(Source – DCC)

3.3.10 1999 Aerial Photograph

In the 1999 aerial photograph, it appears that the landfill footprint has expanded to include most, if not all, of the current footprint.

The weighbridge has been built in its current location along the main access road to the site.

Active filling is concentrated at the southeast corner and along the central southern boundary of the site. Revegetation is ongoing across much of the site (including the central, eastern and western portions) indicating that active filling is not being undertaken in these areas for some time. Wind-break planting has been established on the northern and western boundary of the landfill.

There appears to be a landfill gas collection network under construction or use in the northern and part of the western portion of the site.

(Source – DCC)

3.3.11 2013 Aerial Photograph

By 2013, the Site is generally in its current configuration. The resource recovery facility / transfer station, wind break planting, the access track (and interceptor trench), and the compost facility are all visible in this aerial photograph. Landfilling appears to be active in the northern portion of the Site. The area of capping completed in 2009 / 2010 is noticeable as vegetated (green) in the southeast corner of the landfill, as is an area of capping underway at that time immediately adjacent to the west.

(Source – Retrolens)

3.4 Landfilling extent

A review of the available historical aerial photographs and leachate collection trench construction report indicates that waste is present on the outside of the leachate collection trench in places, with waste up to 6 m thick on the north eastern boundary in a short section (approximate 20-40 m length). Generally, the thickness of the waste is less than 2.6 m in other areas, and averaging 1.63 m depth over the other 40 locations mapped from the log.

The aerial photographs (1978 and 1982) indicate that landfilling activities likely extended very close to the edge of the Kaikorai Stream banks along the northern boundary with some waste placement also likely to having occurred in these areas. The likely extent of the landfilling is shown on Figure 4, overleaf.

Data source: World Imagery: LINZ Hybrid Reference Layer: LINZ, Stats NZ, Esri, HERE, Garmin, Foursquare, METI/NASA, USGS. Created by: jholloway

4. Site works

Field work was carried out in compliance with the project specific Job Safety and Environmental Analysis (JSEA) plan. All site investigation staff were inducted to the Site by Waste Management Ltd, the operators of the landfill. Environmental sampling was undertaken between 1 November and 9 November 2022, comprising the following:

- A GHD environmental scientist was on site to assess and sample soil from eight boreholes installed as part of the geotechnical investigation (BH-100 through BH-104, BH-107, BH-108 and BH-111).
- A GHD environmental scientist was on site for the installation of monitoring wells in boreholes BH100, BH-101, and BH-104.
- Soil samples from boreholes BH-102, BH-103, BH-107, BH-108 and BH-111 were collected from core boxes by a GHD environmental scientist. The remainder of the samples were collected during the drilling of BH-100, BH-101 and BH-104.
- Collection of 26 original and 2 duplicate soil samples from the above-mentioned boreholes. Samples were collected from multiple depths in each borehole. The sample depths were chosen based on an assessment of the soil by a GHD environmental scientist.

A site investigation points location plan is presented as Figure 5, overleaf.

As the Site is a known piece of contaminated land, a land use consent was required to drill and install the boreholes. Resource consent RM21.467 (to drill 17 investigation sites on contaminated land for the purpose of geotechnical investigation) was granted by ORC to DCC to undertake these works on 10th October 2021.

Prior to breaking ground, subsurface utilities clearance was undertaken by a specialist contractor (Fulton Hogan), using both a Cable Avoidance Tool (CAT) and Ground Penetrating Radar (GRP) at all borehole locations. A Ground Penetration Permit to Work was issued by an authorised GHD permitter.

Logging of the material encountered during drilling was undertaken by a GHD geotechnical engineer in accordance with NZ Geotechnical Society (2005) field description of soil and rock. Borehole logs are included in Attachment B.

Data source: World Imagery: LINZ Hybrid Reference Layer: LINZ, Stats NZ, Esri, HERE, Garmin, Foursquare, METI/NASA, USGS. Created by: jholloway

4.1 Soil sampling methodology

The geotechnical investigation was mainly focussed on the peripheries of the landfill to better understand liquefaction potential of the material present in these areas. Consequently, the environmental investigation was limited to the assessment of the material collected during the drilling works in these areas.

Environmental soil samples were collected from a total of eight boreholes at multiple depths at each borehole to target different depths and lithologies to gain an understanding of the material present. GHD collected two duplicate samples for quality control and assurance purposes.

Soil samples were collected in accordance with standard GHD procedures. Samples were placed directly into laboratory supplied containers and then placed in an iced chilly bin and couriered to ALS - Analytica Laboratories under standard GHD Chain of Custody (CoC) procedures. Due to capacity issues at Analytica Laboratories, a portion of the polycyclic aromatic hydrocarbon and semivolatile organic compounds analysis was subcontracted to Eurofins Environmental Testing NZ Limited (Eurofins).

A total of 26 primary soil samples were selected for analysis. Based on the historical land use and observations made during sampling, selected soil samples were analysed for the identified contaminants of concern as follows:

- Heavy metal suite (arsenic, beryllium, boron, cadmium, chromium, copper, lead, mercury, nickel and zinc) x26 samples
- Polycyclic Aromatic Hydrocarbons (PAHs) x23 samples
- Semi Volatile Organic Compounds (SVOCs) x13 samples
- Total Petroleum Hydrocarbons (TPH) x1 sample
- Asbestos in soil (presence/absence) x18 samples
- Asbestos in soil (semi-quantitative) x1 sample
- Ammonia x8 samples
- Two duplicate samples, one of which was analysed for heavy metals and PAHs and the other was analysed for heavy metals only.

4.2 Field Observations

Solid waste and/or fill material was encountered in all eight of the boreholes that were sampled as part of this environmental assessment.

Boreholes BH-100 to BH-104, which were installed on the perimeter of the known landfill area, contained fill and/or solid waste to depths up to 3.95 metres below ground level (m bgl). Boreholes BH-107, BH-108 and BH-111 were installed on the landfill and contained fill to at least 7.5 m bgl, 11.2 m bgl and 4.6 m bgl respectively.

A summary of the borehole locations and the thickness of fill is provided in Table 4 below. Further details can be found on the borehole logs in Attachment B.

Number borehole	(m bgl)	Borehole Depth of Thickness of Fill / Waste (m)	Location
BH-100	12.95	2.75	Inside / within the trench
BH-101	12.95	1.3	Outside the trench
BH-102	14.95	1.6	Inside the trench (Bund slope)
BH-103	13.15	$1.2 - 1.5$ (Core loss)	Outside the trench
BH-104	9.95	$3.95 - 4.5$ (Core loss)	Inside the trench

Table 4: Borehole locations and fill thicknesses summary

A Photo-ionisation detector (PID) was used to screen soil samples for volatile contaminants at all locations, with the exception of borehole BH-100. PID measurements are included in Table 6. It should be noted that some of the PID readings were collected during drilling works while the others were measured during the collection of soil samples from core boxes. For the samples that were collected from the core boxes, the measured PID values are likely lower than what would have been measured if the samples were collected during the drilling works, as the loss of volatiles would likely have occurred over the storage interval.

Note: ppm VOC – parts per million Volatile Organic Compound

5. Regulatory Context

5.1 Applicable Soil Contaminant Standards

5.1.1 The National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health

The User's Guide: National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health (MfE, 2012)⁴ details Soil Contaminant Standards (SCSs) for seven inorganic substances and five organic compounds (or groups of compounds). SCSs are available for these substances and compounds when present in land used for five land use scenarios. The contaminants analysed at this site for which SCSs are available are arsenic, cadmium, chromium, copper, lead, mercury, DDT (Dichlorodiphenyltrichloroethane) and benzo(a)pyrene equivalent (BaP). The NESCS applies to a "piece of land" on which a HAIL activity has occurred or is currently occurring.

The land use category selected for the purposes of this investigation was Commercial / Industrial and is described in the NESCS User Guide as "Commercial / Industrial site with varying degrees of exposed soil. Exposure of outdoor workers to near-surface soil during routine maintenance and gardening activities with occasional excavation as part of maintaining subsurface utilities. Also, conservatively applicable to outdoor workers on a largely unpaved site".

Because the landfill area may, at some point in the future, be used for recreational purposes (eg.as a reserve or park), a Recreational land use scenario was also considered. The NESCS User Guide describes Recreational land use as "Public and private green areas and reserves used for active sports and recreation. This scenario is intended to cover playing fields and suburban reserves where children play frequently. It can also reasonably cover secondary school playing fields but not primary school playing fields."

These land-use exposure scenarios have been adopted for screening purposes to include potential receptors including site workers during any future construction works, current and future users of the Site and ongoing maintenance / excavation workers at the site.

The intention of the NESCS is the protection of human health from contaminated land and the appropriate assessment of the risk to human health prior to the undertaking of the regulated activities (e.g. Soil disturbance or land use change).

If the investigation demonstrates that the contaminants tested are at, or below, screening criteria concentrations, the regulations of the NESCS will not apply should any of the regulated activities be undertaken over the Site.

NESCS SCS criteria adopted for the Site are presented in Table 1 through to Table 3 in Attachment C.

5.1.2 Health and Safety at Work (Asbestos) Regulations 2016

The management and/or removal of asbestos in soils is regulated under the Health and Safety at Work (Asbestos) Regulations 2016⁵ (Asbestos Regulations). However, the Asbestos Regulations do not provide guidance regarding the definitions of what constitutes an asbestos contaminated site, in particular, with regard to soil. Rather, the Regulations simply states that the Asbestos Regulations apply where a competent person advises that the disturbance and/or removal of soil is likely to lead to airborne contamination at a level that exceeds trace concentrations.

⁴ Ministry for the Environment, 2012. Users Guide: National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health. Wellington: Ministry for the Environment

 5 Health and Safety at Work (Asbestos) Regulations 2016 (15 February 2016) made under sections 24(1) (m), 211, and 218 of the Health and Safety at Work Act 2015

The New Zealand Guidelines for Assessing and Managing Asbestos in Soil (BRANZ Guidelines 2017)

The BRANZ Guidelines 2017 \degree provide a methodology to ensure that management of asbestos in soil meets regulatory requirements and an acceptable level of managed risk. This methodology is consistent with the MfE CLMG's for New Zealand and the NESCS for the assessment of asbestos in soil.

If asbestos is detected in soils using a laboratory presence/absence test, the BRANZ Guidelines 2017 then provides an additional guidance for further soil sampling and criteria for the definition of whether the removal of such material is considered licenced asbestos removal (Class A or B), asbestos-related-works or unlicensed works under the Asbestos Regulations.

The adopted asbestos criteria are presented in Table 6 in Attachment C.

5.2 Other applicable Human Health Standards

For contaminants of concern that are not listed as priority contaminants, the NESCS references the Contaminated Land Management Guidelines No. 2: Hierarchy and Application in New Zealand of Environmental Guideline Values (CLMG No.2) to provide guidance.

In the absence of New Zealand risk based human health criteria for certain contaminants of concern, such as nickel and zinc, the Australian National Environment Protection Measure 2013⁷ (NEPM) Recreational (C) and Commercial/Industrial (D) guideline values have been adopted for this investigation.

For TPH, naphthalene and pyrene, Tier 1 screening criteria from the MfE Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand⁸ have been selected. Tier 1 soil acceptance criteria for Commercial / Industrial land use and Residential land use, All Pathways, Sand 1 - 4 m depth have been adopted. As there are no recreational land use guideline values available, the residential values have been adopted as a conservative screen for this pathway.

The adopted criteria are presented in Tables 1 through to 4 in Attachment C.

5.3 Background Soil Concentrations

Background soil concentrations of heavy metals for the Site were obtained from the Landcare Research (2015) report⁹, Tables 13 through Table 15. The area of the landfill is classified as fill and as such does not have background heavy metal concentrations derived. As material has been received at the Site from all over Dunedin, the highest background value from each of the dominant adjacent geological units for each metal was adopted for comparison purposes.

There are currently no published background soil concentrations for PAHs in Dunedin. As such, PAH values established for Christchurch urban soils¹⁰ were adopted for comparative purposes only.

Background concentrations for OCP pesticides were taken from the following two documents:

- Landcare Research (2015). Background soil concentrations of selected trace elements and organic contaminants in New Zealand. Table 21.
- Ministry for the Environment (1998). Ambient concentrations of selected organochlorines in soils. Table F5.1 mean values.

The adopted criteria are presented in Table 1, Table 2 and Table 3 in Attahment C.

 6 New Zealand Guidelines for Assessing and Managing Asbestos in Soil - Building Research Association of New Zealand (BRANZ), November 2017

⁷ National Environmental Protection Council (NEPC) (2013). National Environmental Protection Measure (Assessment of Site Contamination) as amended in 2013 Schedule B1, Health Investigation Levels (HIL) for soil contaminants.

 8 Ministry for the Environment (1999, revised 2011). Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4- Tier 1 Soil screening Criteria.

⁹ Landcare Research (2015). Background soil concentrations of selected trace elements and organic contaminants in New Zealand

¹⁰ Environment Canterbury, 2007, Background concentrations of polycyclic aromatic hydrocarbons in Christchurch urban soils. Report R07/19

5.4 Otago Regional Council Regional Plans

The operative Regional Plan for Otago (Waste) was enacted in April 1997. Section 5 of this Plan details the policies, methods and rules regarding Contaminated Sites. A contaminated site is defined as a site at which hazardous substances occur at concentrations above background levels and where assessment indicates it poses, or is likely to pose, an immediate or long term hazard to human health and/or the environment.

Hazardous substances are defined as follows:

Hazardous substances are substances which impair human, plant or animal health, or which may adversely affect the health or safety of any person or the environment, whether or not they are contained in or form part of any other substance or thing. These include pesticides, petrol, oil, cleaners and paint.

Rule 5.6.1 of the Plan states the following:

Rule 5.6.1 Hazardous wastes at contaminated sites (discretionary activity)

- 1. The disturbance of land; or
- 2. The discharge of hazardous waste into water; or
- 3. The discharge of hazardous waste onto or into land in circumstances that may result in that hazardous waste (or any other hazardous waste emanating as a result of natural processes from that hazardous waste) entering water; or
- 4. The deposit of any hazardous waste, in, on or under land; or
- 5. The discharge of hazardous waste into air at or from a contaminated site;

is a discretionary activity.

The operative Regional Plan for Otago (Water for Otago) was enacted in February 1998 and updated in June 2021. Section 7 of the Regional Plan, details policies for the discharge of stormwater and the Rules associated with these policies are contained in Section 12 (water takes, use and management). However, many of the Rules related to the discharge of stormwater within this Section have been repealed.

Rule 12.B.3 (Discharge of hazardous substances, hazardous wastes, specified contaminants, and stormwater, and discharges from industrial or trade premises and consented dams) states that the discharge of stormwater to water, or onto or into land in circumstances where it may enter water, is a restricted discretionary activity. The restrictions include the potential for soil contamination.

Should contaminants be found present in the soil at the Site at concentrations that there is a hazard posed to human health and/or the environment, then a consent may be required under the Regional Council's rules for certain activities to be undertaken. Further discussion is provided in Section 7.3 on consenting requirements.

6. Results

6.1 Soil analytical results

Analytical results compared against the adopted standards and guidelines are presented in Tables 1 through Table 6 in Attachment C. The laboratory reports are included in Attachment D.

In summary:

- Heavy metals
	- Metals including arsenic, cadmium, chromium, lead, nickel and zinc were detected at concentrations above the adopted Dunedin background soil concentrations in 13 of the 28 samples analysed.
	- No heavy metals were reported at concentrations above the adopted background values in samples collected from BH-100 and BH-107.
	- Above background concentrations were limited to the upper 2.5 m of material in boreholes BH-101, BH-102 and BH-103. However, they extended to at least 3.3 bgl in boreholes BH-108 and BH-111.
	- No metals concentrations were detected above the relevant human health screening criteria.
	- Refer to Table 1 for tabulated analytical metals results.
- PAHs
	- Various PAHs were detected above adopted background soil concentrations in eight samples (not including the field duplicate sample).
	- No PAHs were reported at concentrations above the adopted background values in samples collected from BH-100 and BH-101, BH-102 and BH-103.
	- Above background concentrations of PAHs were distributed through the whole soil profile (surface to 3.3 m bgl).
	- PAH, as BaP toxicity equivalence quotient (TEQ), was greater than the NESCS SCS for Commercial/Industrial and Recreational use in one sample (12547621-BH-104(0.5)) collected at borehole BH-104 in fill material.
	- Refer to Table 2 for the tabulated PAH results.
- SVOCs
	- Various SVOC were detected above background soil concentrations in five samples, two collected from BH-103 (0.2 and 1.2 m bgl), two collected from BH-107 (1.5 and 2.5 m bgl) and the other from BH-111 (1.5 m bgl).
	- The pesticide 4,4'-DDD was reported present at a low concentration in boreholes BH-103, BH-107 and BH111 in the samples described above. The pesticides 4,4'-DDE and 4,4'-DDT were also reported present in the samples collected and analysed from BH-103.
	- No SVOCs were detected at concentrations above any considered human health or environmental screening criteria.
	- Refer to Table 3 for a summary of SVOC results. Only those SVOCs which were detected at concentrations above the LOR are presented in the table. The full set of results can be found in the laboratory reports.
- TPH was not detected above any of the considered human health or environmental screening standards. Refer to Table 4 for tabulated TPH results.
- Ammonia was detected above the laboratory detection limit in four of eight samples. The highest concentrations were found in the samples collected from soils likely to be rich in organic material (on the basis of their lithological description). Refer to Table 5 for a summary of ammonia results. There are no available guideline criteria available to compare these results against.

– Asbestos

- Asbestos fines were reported present in one sample collected from BH-107 at a depth of 0.5 m bgl (12547621-BH-107(0.5)), and the sample was further scheduled for semi-quantitative analysis. Asbestos was not detected in the sample examined for semi-quantitative analysis.
- Asbestos was not detected in any other soil samples collected as part of this investigation.
- Refer to Table 6 for a summary of asbestos results.

6.2 Quality Assurance and Quality Control

GHD quality assurance/quality control (QA/QC) procedures to assess data quality were maintained throughout the project.

The QA/QC programme undertaken as part of the assessment by GHD included the following:

- All fieldwork was undertaken by suitably qualified and trained staff. The work managed by a suitably qualified and experienced practitioner (SQEP) and the report was reviewed by a SQEP, as required by the NESCS.
- Collection of two duplicate soil samples for analysis of heavy metals and PAHs.
- Soil samples collected throughout works were dispatched to Analytica Laboratories in Hamilton on the day of collection under standard chain of custody procedures and the analysis was undertaken within the sample holding times.
- Analytica Laboratories is an internationally recognised laboratory endorsed by International Accreditation New Zealand (IANZ) which represents New Zealand in the International Laboratory Accreditation Cooperation (ILAC). Through the ILAC mutual recognition arrangement (ILAC-MRA) this accreditation is internationally recognised. The tests were performed in accordance with the terms of accreditation.

6.2.1 Analytical results

GHD collected two field duplicate samples for quality assurance/quality control (QA/QC) one of which was analysed for both heavy metals and PAHs and the other for heavy metals only.

A quantitative measure of the accuracy of the analytical results received from the laboratory was conducted using calculated relative percentage difference (RPD) values. The RPD values were calculated using the following equation:

RPD (%) $= \langle Co - Cs \rangle$ $\overline{Co + Cs \setminus}$ $\overline{}$ $^{+}$ 2 x 100

Where $Co =$ concentration obtained from the original sample. Cs = concentration obtained from the duplicate sample.

The usual acceptance criteria for relative percentage difference (RPD) are between 30 and 50% in soils (CLMG, No.5 section 3.9.1 (2011)). However, larger RPDs are allowed for different analytes (up to 100%). A large percentage differential can occur particularly in soils due to the following:

- A small analytical differential between two samples based on the low levels of detection from the primary and duplicate soil sample.
- Soil samples collected from a non-homogenous (heterogeneous) soil profile.

RPD values for the duplicate samples (field) analysed for heavy metals for this assessment ranged from 0% to 61.6%. However, the majority of the RPDs, 17 out of 20, were less than 50%.

RPD values for the duplicate samples (field) analysed for PAHs for this assessment ranged from 127.5% to 187.9%.

The high RPD values for the PAH analysis are likely due to the heterogenous nature of the material at that borehole location (BH-108).

The duplicate quality assurance and quality control (QA/QC) results are provided in Attachment C, Table 7 and the laboratory analytical reports are provided in Attachment D.

The results of the whole QA/QC program are considered to provide an acceptable degree of confidence in the sampling and analytical program.

7. Conclusions

7.1 Landfilling Extent

A review of the available historical aerial photographs and leachate collection trench construction report indicates that waste is present on the outside of the leachate collection trench in places, with waste up to 6 m thick on the north eastern boundary in a short section (approximate 20-40 m length). Generally, the thickness of the waste is less than 2.6 m in other areas, and averaging 1.63 m depth over the other 40 locations mapped from the log.

The aerial photographs (1978 and 1982) indicate that landfilling activities likely extended very close to the edge of the Kaikorai Stream banks along the northern boundary with some waste placement also likely to having occurred in these areas. The trench construction report indicates that waste / landfill material is present on the outside of the trench at the following locations:

- 6m north of PS3 through to 32m east of AV4
- 36m east of PS5 to 49m east of PS5
- MH6 through to 18m north of PS8

Waste is likely to extend either side of these locations. In addition, earthfill is also present from south of PS8 through to the end of the trench (55 m north of PS9).

The likely extent of the landfilling is shown on Figure 4.

7.2 Soil contaminant status

An environmental site investigation was undertaken in conjunction with a geotechnical investigation in October and November 2022. Soil samples were collected from eight of the boreholes drilled during the investigation, six of which were located around the periphery of the landfill and two in the northern portion of the landfill (area of oldest landfilling). These samples were laboratory analysed for the identified contaminants of concern. The material in the boreholes was logged by a geotechnical engineer and observations of soil type and indications of contamination were also made by an environmental scientist.

Municipal solid waste (MSW) and fill material was found present in all the boreholes assessed.

The analytical results indicated that contaminant concentrations at above background values were present through the soil profile from surface to a depth of 3.3 m bgl. One sample, at BH-104(0.5 m bgl), had a reported a BaP equivalent concentration which exceeded the NESCS SCS value for both Commercial / Industrial and Recreational land use.

Asbestos fines were reported present in only one of the 19 samples analysed. Low concentrations of TPH were reported present in the one sample analysed for this contaminant. Elevated concentrations of ammonia were reported in samples collected from borehole horizons described as peat, silty sand, fill / organic silt and fill / wood in sand. These materials are likely to have elevated organic content.

Pesticides (4,4'-DDE, 4,4'-DDD and 4,4'-DDT), at low concentrations, were reported present in five of the thirteen samples analysed for these contaminants.

7.3 Regulatory requirements

7.3.1 NESCS

This investigation has confirmed that HAIL activity G3 (landfill sites) has occurred on the Site and as such the requirements of the NESCS apply to any future redevelopment works (e.g. soil disturbance and land use change).

7.3.2 Regional and Local Council Requirements

It is understood that a review of Regional and Local Council planning requirements is to be undertaken by Boffa Miskell. As such, further discussion of any requirements has not been undertaken in this report.

Attachments

Attachment A Historical Plans and Figures

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Attachment B Borehole Logs

GLOSSARY OF SYMBOLS

This standard sheet should be read in conjunction with all test hole log sheets and any idealised geological sections prepared for the investigation report.

GENERAL ABBREVIATIONS

PT Pressuremeter Test
R Rising Head Permea

R Rising Head Permeability Test
SV Shear Vane Test (suffixed by v

SV Shear Vane Test (suffixed by value in kPa, peak/residual values)
UTP Unable to penetrate (shear vane testing)

- UTP Unable to penetrate (shear vane testing)

TD Target depth
- Target depth
- HCL Hydrochloric acid

WELL SYMBOLS

GROUNDWATER SYMBOLS

Groundwater level

\\ghdnet\ghd\NZ\Manukau\Projects\51\0101204 Geotech\08 Geotech Field Resources and Calibration\GHD standard BH sheets\Descriptions\ss\NZStandardSheet_SymbolGlossary_ver7.doc 1

SOIL SYMBOLS

Main Components

Note: Composite soil types will be signified by combined symbols, e.g $\left| \cdot \right|$ Sandy CLAY

SOIL DESCRIPTION ABBREVIATIONS

Moisture Condition

ROCK SYMBOLS

Note: Additional rock symbols may be allocated for a particular project. Interbedded rock will be represented using alternatively the above symbols

\\ghdnet\ghd\NZ\Manukau\Projects\51\0101204 Geotech\08 Geotech Field Resources and Calibration\GHD standard BH sheets\Descriptions\ss\NZStandardSheet_SymbolGlossary_ver7.doc 2

ROCK DESCRIPTION ABBREVIATIONS

DEFECT DESCRIPTION ABBREVIATIONS

Fracture Type
BP Bedding

- BP Bedding Plane
CB Cross Bed
- CB Cross Bed
Cl Cleavage
- Cl Cleavage
CS Crushed
- CS Crushed Seam
CZ Crush zone
- CZ Crush zone
Fl Foliation
- Fl Foliation
FZ Fracture
- FZ Fractured Zone (>250 mm)
JS Joint set
- Joint set

Inclination
SB Sub-

- SB Sub-horizontal
G Gently inclined
- G Gently inclined
M Moderately incl
- M Moderately inclined
S Steeply inclined
- S Steeply inclined
VS Verv steeply inc
- VS Very steeply inclined
SV Sub-vertical
- Sub-vertical

Roughness

- Slickensided
- r Rough
- sm Smooth

Texture

- Pl Planar
St Steppe
- St Stepped
U Undulatii
- Undulating

Joint Set Counts

- X 2 2 joints
X 3 3 joints
- X 3 $\overline{)}$ 3 joints
 X 4 $\overline{)}$ 4 joints
- X 4 4 joints
 X 5 5 joints
- X 5 5 joints
X 6 6 joints
- X 6 6 joints
X 7 7 joints 7 joints
- X 8 8 joints
- X 9 9 joints
- > 10 > 10 joints

Core Recovery Parameters

- TCR Total Core Recovery %
- SCR Solid Core Recovery % RQD – Rock Quality Designation % 97 100 100
	-

Weathering

-
- JT Joint SF Sheared Surface
SM Seam
- SM Seam
SS Shear
- SS Sheared Seam
SZ Sheared Zone (
- SZ Sheared Zone (>250 mm)
VN Vein
- Vein

Aperture
 T Tig

- T Tight
VN Verv
- Very Narrow
- N Narrow
MN Modera
- MN Moderately Narrow
MW Moderately Wide
- MW Moderately Wide
W Wide
- Wide VW Very Wide

Infilling or Coating

- CN Clean
X Carbo
- Carbonaceous
- CLAY Clay
- KT Chlorite
CA Calcite
- Calcite
- Fe Iron Oxide
- MI Micaceous
OZ Ouartz
- QZ Quartz
VE Veneer
- Veneer

Spacing

- EC Extremely closely spaced
VC Very closely spaced
- VC Very closely spaced
C Closely spaced
- C Closely spaced
MW Moderately wide
- MW Moderately widely spaced
W Widely spaced
- W Widely spaced
VW Very widely spa
- Very widely spaced

Visual Defects

Visual representation of defect angle from horizontal (example shown is 45°)

Report ID: GENERAL_LOG || Project: GILF LOGS 24112022.GPJ || Library: GHD - NZGD.GLB || Date: 13 December 2022

Attachment C Analytical Results Tables

3. Landcare Research Limited (2006) PBC - Predicted Background Soil Concentrations, New Zealand - fill material (the highest value for each metal from the dominant soil groups surrounding Green Island Landfill) Dunedin , N 4. NESCS SCS criteria presented are for Chromium (VI)

Notes:

Values shaded grey exceed the NESCS SCSs for Protection of Human Health (Commercial / Industrial land use) Values shaded yellow exceed the NESCS SCSs for Protection of Human Health (Recreational land use) Red Text exceed the adopted Dunedin Background Soil Concentrations (mg/kg) SCS ‐ Soil contaminant standard

‐ Data not available or not analysed in this sample All units are in mg/kg m bgl ‐ metres below ground level

* Field duplicate of 12547621‐BH‐108(1.9)

** Field duplicate of 12547621‐BH‐111(3.2)

References

1. Ministry for the Environment (2011). Resource Management (National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health) Regulations 2011 (NESCS).

2. National Environment Protection Council (1999, revised 2013) National Environment Protection (Assessment of Site Contamination) Measure. Table 1A. (NEPM) Health Investigation Levels Commercial / Industrial D.

Site Name: Green Island Landfill Table 2: PAH Analytical Results November 2022 Site Investigation

g/kg)

Values shaded according according to the highest exceedance
All units are in mg/kg
m bgl - metres below ground level
A hyphen (-) indicates criterion not available or sample not anlaysed for this analyte
A - reported at a

Limiting pathways: (v) ⁼ volatilisation, (p) ⁼ Produce (2) ⁼ Brackets denote values exceed threshold likely to correspond to formation of residual separate phase hydrocarbons

NA - indicates contaminant not limiting as estimated health‐based criterion is significantly higher than that likely to be encountered on site.

References
1. Ministry for the Environment (2011). Resource Management (National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health) Regulations 2011 (NESCS). Soil Contaminant St

2. Ministry for the Environment (2011) Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4 - Tier 1 Soil screening criteria. Table 4.10. Residential land use and Table 4.

3. Environment Canterbury (2007). Background Concentrations of polycyclic aromatic hydrocarbons in Christchurch urban soils. Report No. R07/19. Table 9.
4. Benzo(a)pyrene equivalent concentration calculated as the sum of t

Site Name: Green Island Landfill Table 3: Semi volatile organic compounds (SVOC) Analytical Results November 2022 SIte Investigation

Notes:

 Values shaded grey exceed the NESCS SCSs for Protection of Human Health (Commercial / Industrial land use) Values shaded yellow exceed the NESCS SCSs for Protection of Human Health (Recreational land use) Value shaded peach exceed the MfE Hydrocarbon Guidelines for Residential land use Value shaded green exceed the MfE Hydrocarbon Guidelines for Commercial / Industrial land use

Red text indicates values exceeds the adopted background concentration value

mg/kg ‐ miligrams per kilogram

m bgl ‐ metres below ground level

A hyphen (‐) indicates criterion not available for this analyte

< ‐ reported at ^a concentration less than the laboratory limit of reporting (LOR)

References:

1. Resource Management (National Environmental Standard for Assessing and Managing Contaminants in Soil to Protect Human Health) Regulations 2011 (NESCS). Soil contaminant standards (SCS) for Commercial / Industrial Land use

2. Ministry for the Environment (2011) Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4 - Tier 1 Soil screening criteria. Table 4.10. Residential land use and Table 4.

3. Benzo(a)pyrene equivalent concentration calculated as the sum of the carcinogenic PAHs in accordance with the methodology published in the NESCS.

3. US EPA Superfund EGV (June, 2013)

4. Ministry for the Environment (1998). Ambient concentrations of selected organochlorines in soils. Table F3 ‐ mean values.

5. Environment Canterbury (2007). Background Concentrations of polycyclic aromatic hydrocarbons in Christchurch urban soils. Report No. R07/19. Table 9.

6. Landcare Research Limited (2015) Background soil concentrations of selected trace elements and organic contaminants in New Zealand . Table 21. Christchurch Soils.

Notes:

All units are in mg/kg

m bgl ‐ metres below ground level

A hyphen (‐) indicates criterion not available

< ‐ reported at ^a concentration less than the laboratory limit of reporting (LOR)

Limiting pathways: (m) = Maintenance/Excavation , (2) = Brackets denote values exceed threshold likely to correspond to formation of residual separate phasehydrocarbons, (x) = PAH surrogate NA ‐ indicates estimated criterion exceeds 20,000 mg/kg. At 20,000 mg/kg residual separate phase is expected to have formed in soil matrix. Some aesthetic impact may be noted.

References

1. Ministry for the Environment (2011) Guidelines for Assessing and Managing Petroleum Hydrocarbon Contaminated Sites in New Zealand. Module 4 ‐ Tier 1 Soil screening criteria. Table 4.14. Commercial / Industrial land use.

Green Island Landfill Soil Analytical Results November 2022 Table 5: Ammonia in Soil

Notes:

All units are in mg/kg m bgl ‐ metres below ground level

Green Island Landfill Soil Analytical Results November 2022 Table 6: Asbestos in Soil

Notes:

Peach shading indicates that asbestos was found present but at concentrations less than the guideline value

Blue shading indicates asbestos present at concentrations above the adopted BRANZ guideline values

* This sample was ^a retest of the sample collected at the same location and depth on 1 November 2022

Reference:

1. BRANZ (2017) New Zealand Guidelines for Assessing and Managing Asbestos in Soil. Table 5 ‐ Commercial / Industrial guideline values

Dunedin City Council Green Island Landfill Soil Sampling Results 12547621

Attachment D Laboratory Reports

Analytica Laboratories Limited Ruakura Research Centre 10 Bisley Road Hamilton sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

9 Heavy Metals in Soil

9 Heavy Metals in Soil

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited.

This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.

9 Heavy Metals in Soil

9 Heavy Metals in Soil

9 Heavy Metals in Soil

Soil Aggregate Properties and Nutrients

Soil Aggregate Properties and Nutrients

Custom Job

Custom Job

Custom Job

Method Summary

Elements in Soil Samples dried and passed through a 2 mm sieve followed by acid digestion and analysis by ICP-MS. In accordance with in-house procedure based on US EPA method 200.8.

Ammonia-N in Soil1:5 water extraction (NEPM, Schedule B3, Laboratory Analysis of Potentially Contaminated Soil, 2011) followed by colour-metric analysis (APHA 4500 NH₃ - F Online edition - modified - Discrete Analyser).

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Jarred Wilson, DipSci Trace Elements Team Leader

Sandra Mathews, B.Eng. Technologist

Adam Ang **Team Leader**

Sharelle Frank, B.Sc. (Tech) Technologist

Analytica Laboratories Limited 34 Brisbane Street Sydenham **Christchurch** sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

Sampling Site: Green Island Landfill Description of Work: PA - 12547621

Lab Reference: 22-40290 Submitted by: Danny Fitzgerald Date Received: 04/11/2022 Testing Initiated: 7/11/2022 Date Completed: 8/11/2022 Order Number: Reference: 12547621

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

Asbestos in Soil (Qualitative) Sample Details

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled.

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited. This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.

Information in the above table supplied by the client: Client Sample ID.

Asbestos in Soil (Qualitative) Approver:

Av. Eeden

Aleesha van Eeden, M.Sc. Technician

Method Summary

Asbestos Fibres in Soil (Qualitative)

Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in bulk samples.

Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is not the sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.

Analytica Laboratories Limited 34 Brisbane Street Sydenham **Christchurch** sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

Sampling Site: Green Island LF Description of Work: Soil - Green Island LF

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

AMENDED REPORT. This report replaces in full a previous version 22-40546[R00] sent on 09/11/2022. Changes have been made to the Client Sample ID's as per request of the client.

Asbestos in Soil (Qualitative)

Sample Details

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled.

Information in the above table supplied by the client: Client Sample ID.

Asbestos in Soil (Qualitative) Approver:

ceden

Aleesha van Eeden, M.Sc. Technician

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\kappa_c c^{\alpha EDI}r_{\xi_0}$ marked *, which are not accredited.

Method Summary

Asbestos Fibres in Soil (Qualitative) Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in bulk samples.

> Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is not the sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.

Analytica Laboratories Limited Ruakura Research Centre 10 Bisley Road Hamilton sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

Sampling Site:

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report.

Specific testing dates are available on request.

Heavy Metals in Soil

Polycyclic Aromatic Hydrocarbons - Soil

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited. This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.

Polycyclic Aromatic Hydrocarbons - Soil

Total Petroleum Hydrocarbons - Soil

Moisture Content

Semivolatile Organic Compounds - Soil

Semivolatile Organic Compounds - Soil

Semivolatile Organic Compounds - Soil

Method Summary

Elements in Soil Samples dried and passed through a 2 mm sieve followed by acid digestion and analysis by ICP-MS. In accordance with in-house procedure based on US EPA method 200.8.

PAH in Soil Solvent extraction, silica cleanup, followed by GC-MS analysis. **Benzo[a]pyrene TEQ (LOR)**: The most conservative TEQ estimate, where a result is reported as less than the limit of reporting (LOR) the LOR value is used to calculate the TEQ for that PAH. **Benzo[a]pyrene TEQ (Zero)**: The least conservative TEQ estimate, PAHs reported as less than the limit of reporting (LOR) are not included in the TEQ calculation. Benzo[a]pyrene toxic equivalence (TEQ) is calculated according to *'Methodology for Deriving Standards for Contaminants in Soil to Protect Human Health'*. Ministry for the Environment. 2011. (In accordance with in-house procedure).

TPH in Soil Solvent extraction, silica cleanup, followed by GC-FID analysis. (C7-C36). (In accordance with inhouse procedure based on US EPA 8015).

Moisture Moisture content is determined gravimetrically by drying at 103 °C.

Astra Southwood

Lab Technician

SVOC in Soil Solvent extraction, followed by GC-MS analysis.(In-house based on US EPA 8270).

Thara Samarasinghe, B.Sc. Technician

Prianshu Chawla, B.Tech Technologist

Jukoubo

Yuri Zubenko, Ph.D. Senior Technologist

Analytica Laboratories Limited Unit 1, 30 Greenpark Road Penrose Auckland sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd Level 1, Bing Harris Building, 286 Princess Street, Dunedin Submitted by: Dunedin 9016 Attention: Danny Fitzgerald Phone: 021 973 994 Email: danny.fitzgerald@ghd.com

Sampling Site: Green Island LF Description of Work: Soils - Green Island LF

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

Lab Reference: 22-40968

Date Received: 09/11/2022
Testing Initiated: 9/11/2022

Date Completed: 14/11/2022

Testing Initiated:

Order Number: N/A Reference: 12547621

Danny Fitzgerald

Asbestos in Soil (Semi-Quantitative)

Sample Details

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled

Analysis Results (Summary)

Information in the above table supplied by the client: Client Sample ID

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited. This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.

Analysis Results (Size Fraction Breakdown)

Information in the above table supplied by the client: Client Sample ID

Asbestos in Soil (Semi-Quantitative) Approver:

Emily Wang, M.Sc. Laboratory Technician

Method Summary

Asbestos Fibres in Soil (Semi-Quantitative)

Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in soil samples.

Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is notthe sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.

Analytica Laboratories Limited Unit 1, 30 Greenpark Road Penrose Auckland sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

GHD Ltd Level 1, Bing Harris Building, 286 Princess Street, Dunedin Submitted by: Dunedin 9016 Attention: Danny Fitzgerald Phone: 021 973 994 Email: danny.fitzgerald@ghd.com

Sampling Site: Green Island LF Description of Work: Soils - Green Island LF

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

Lab Reference: 22-40968

Date Received: 09/11/2022
Testing Initiated: 9/11/2022

Date Completed: 14/11/2022

Testing Initiated:

Order Number: N/A Reference: 12547621

Danny Fitzgerald

Asbestos in Soil (Qualitative) Sample Details

Information in the above table supplied by the client: Client Sample ID, Sample Location, Date Sampled.

Information in the above table supplied by the client: Client Sample ID.

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited. This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.

Asbestos in Soil (Qualitative) Approver:

Emily Wang, M.Sc. Laboratory Technician

Method Summary

Asbestos Fibres in Soil (Qualitative)

Sample analysis was performed using polarised light microscopy with dispersion staining in accordance with AS4964-2004 Method for the qualitative identification of asbestos in bulk samples.

Note 1: The reporting limit for this analysis is 0.1g/kg (0.01%) by application of polarised light microscopy, dispersion staining and trace analysis techniques.

Note 2: Trace asbestos is indicative that freely liberated respirable fibres are present and dust control measures should be implemented or increased on site. This is notthe sole indicator for the friable nature of the asbestos present.

Note 3: If mineral fibres of unknown type are detected, by PLM and dispersion staining, these may or may not be asbestos fibres. To confirm the identity of this fibre, another independent analytical technique such as XRD analysis is advised.

Note 4: The laboratory does not take responsibility for the sampling procedure or accuracy of sample location description.

Analytica Laboratories Limited Ruakura Research Centre 10 Bisley Road Hamilton sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

Heavy Metals in Soil

Heavy Metals in Soil

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited.

This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.

Heavy Metals in Soil

Polycyclic Aromatic Hydrocarbons - Soil

Polycyclic Aromatic Hydrocarbons - Soil

Polycyclic Aromatic Hydrocarbons - Soil

Moisture Content

Moisture Content

Method Summary

PAH in Soil Solvent extraction, silica cleanup, followed by GC-MS analysis. **Benzo[a]pyrene TEQ (LOR)**: The most conservative TEQ estimate, where a result is reported as less than the limit of reporting (LOR) the LOR value is used to calculate the TEQ for that PAH. **Benzo[a]pyrene TEQ (Zero)**: The least conservative TEQ estimate, PAHs reported as less than the limit of reporting (LOR) are not included in the TEQ calculation. Benzo[a]pyrene toxic equivalence (TEQ) is calculated according to *'Methodology for Deriving Standards for Contaminants in Soil to Protect Human Health'*. Ministry for the Environment. 2011. (In accordance with in-house procedure). **Moisture** Moisture content is determined gravimetrically by drying at 103 °C.

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Jarred Wilson, DipSci Trace Elements Team Leader aissocilhenwood Lab Technician

jənshv Prianshu Chawla, B.Tech Technologist

Brent Boynes Lab Technician

Analytica Laboratories Limited Ruakura Research Centre 10 Bisley Road Hamilton sales@analytica.co.nz www.analytica.co.nz

Certificate of Analysis

Report Comments

Samples were collected by yourselves (or your agent) and analysed as received at Analytica Laboratories. Samples were in acceptable condition unless otherwise noted on this report. Specific testing dates are available on request.

Polycyclic Aromatic Hydrocarbons - Soil

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation with the exception of tests $\epsilon_0 e^{cRED/T} \epsilon_0$ marked *, which are not accredited. This test report shall not be reproduced except in full, without the written permission of Analytica Laboratories.

Polycyclic Aromatic Hydrocarbons - Soil

Moisture Content

Method Summary

PAH in Soil Solvent extraction, silica cleanup, followed by GC-MS analysis.

Benzo[a]pyrene TEQ (LOR): The most conservative TEQ estimate, where a result is reported as less than the limit of reporting (LOR) the LOR value is used to calculate the TEQ for that PAH. **Benzo[a]pyrene TEQ (Zero)**: The least conservative TEQ estimate, PAHs reported as less than the limit of reporting (LOR) are not included in the TEQ calculation. Benzo[a]pyrene toxic equivalence (TEQ) is calculated according to *'Methodology for Deriving Standards for Contaminants in Soil to Protect Human Health'*. Ministry for the Environment. 2011. (In accordance with in-house procedure).

Moisture Moisture content is determined gravimetrically by drying at 103 °C.

Prianshu Chawla, B.Tech Technologist

Brent Boynes Lab Technician

Certificate of Analysis

Environment Testing

All tests reported herein have been performed in accordance with the laboratory's scope of accreditation

Analytica Laboratories LTD 10 Bisley Road Hamilton New Zealand 3214

Attention: Customer Service

Report 938712-S Project name Project ID 22-40179 Received Date Nov 08, 2022

Sample History

Where samples are submitted/analysed over several days, the last date of extraction is reported.

If the date and time of sampling are not provided, the Laboratory will not be responsible for compromised results should testing be performed outside the recommended holding time.

Internal Quality Control Review and Glossary

General

- 1. Laboratory QC results for Method Blanks, Duplicates, Matrix Spikes, and Laboratory Control Samples follows guidelines delineated in the National Environment Protection (Assessment of Site Contamination) Measure 1999, as amended May 2013 and are included in this QC report where applicable. Additional QC data may be available on request.
- 2. All soil/sediment/solid results are reported on a dry basis, unless otherwise stated.
- 3. All biota/food results are reported on a wet weight basis on the edible portion, unless otherwise stated.
- 4. Actual LORs are matrix dependant. Quoted LORs may be raised where sample extracts are diluted due to interferences.
- 5. Results are uncorrected for matrix spikes or surrogate recoveries except for PFAS compounds.
- 6. SVOC analysis on waters are performed on homogenised, unfiltered samples, unless noted otherwise.
- 7. Samples were analysed on an 'as received' basis.
- 8. Information identified on this report with blue colour, indicates data provided by customer that may have an impact on the results.
- 9. This report replaces any interim results previously issued.

Holding Times

Please refer to 'Sample Preservation and Container Guide' for holding times (QS3001).

For samples received on the last day of holding time, notification of testing requirements should have been received at least 6 hours prior to sample receipt deadlines as stated on the SRA.

If the Laboratory did not receive the information in the required timeframe, and regardless of any other integrity issues, suitably qualified results may still be reported.

Holding times apply from the date of sampling, therefore compliance to these may be outside the laboratory's control.

For VOCs containing vinyl chloride, styrene and 2-chloroethyl vinyl ether the holding time is 7 days however for all other VOCs such as BTEX or C6-10 TRH then the holding time is 14 days.

Units

Terms

QC - Acceptance Criteria

The acceptance criteria should be used as a guide only and may be different when site specific Sampling Analysis and Quality Plan (SAQP) have been implemented

RPD Duplicates: Global RPD Duplicates Acceptance Criteria is 30% however the following acceptance guidelines are equally applicable:

Results <10 times the LOR: No Limit

Results between 10-20 times the LOR: RPD must lie between 0-50%

Results >20 times the LOR : RPD must lie between 0-30%

NOTE: pH duplicates are reported as a range not as RPD

Surrogate Recoveries: Recoveries must lie between 20-130% for Speciated Phenols & 50-150% for PFAS

PFAS field samples that contain surrogate recoveries in excess of the QC limit designated in QSM 5.4 where no positive PFAS results have been reported have been reviewed and no data was affected.

QC Data General Comments

.

- 1. Where a result is reported as a less than (<), higher than the nominated LOR, this is due to either matrix interference, extract dilution required due to interferences or contaminant levels within the sample, high moisture content or insufficient sample provided.
- 2. Duplicate data shown within this report that states the word "BATCH" is a Batch Duplicate from outside of your sample batch, but within the laboratory sample batch at a 1:10 ratio. The Parent and Duplicate data shown is not data from your samples.
- 3. pH and Free Chlorine analysed in the laboratory Analysis on this test must begin within 30 minutes of sampling. Therefore, laboratory analysis is unlikely to be completed within holding time.Analysis will begin as soon as possible after sample receipt.
- 4. Recovery Data (Spikes & Surrogates) where chromatographic interference does not allow the determination of recovery the term "INT" appears against that analyte.
- 5. For Matrix Spikes and LCS results a dash "-" in the report means that the specific analyte was not added to the QC sample.
- 6. Duplicate RPDs are calculated from raw analytical data thus it is possible to have two sets of data.

Quality Control Results

Environment Testing

Environment Testing

Comments

Qualifier Codes/Comments

- Code Description
- N07 Please note:- These two PAH isomers closely co-elute using the most contemporary analytical methods and both the reported concentration (and the TEQ) apply specifically to the total of the two co-eluting PAHs

Q15 The RPD reported passes Eurofins Environment Testing's QC - Acceptance Criteria as defined in the Internal Quality Control Review and Glossary page of this report.

Authorised by:

Michael Ritchie Senior Analyst-Organic Karishma Patel **Analytical Services Manager**

Michael Ritchie Head of Semi Volatiles (Key Technical Personnel)

Final Report – this report replaces any previously issued Report

- Indicates Not Requested

* Indicates IANZ accreditation does not cover the performance of this service

Measurement uncertainty of test data is available on request or please [click here.](https://cdnmedia.eurofins.com/apac/media/612806/reporting-measurement-uncertainty-of-chemical-and-mycology-test-results-may-2022.pdf)

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E-1 Leachate levels within landfill

Gas wells and groundwater monitoring wells within the landfill were dipped to measure the level of leachate within the landfill footprint (August 2022). These measurements are presented below in Table E.1. Figure E.1 shows the location of the wells. The leachate measurements were undertaken by a site contractor, GHD has not verified these measurements.

Table E.1 Leachate measurements

Dip Dip RL from **Dist from** Top of Level of Level of **Areas and Wells** X (NZTM) Y (NZTM) W/head to W/head Pix 4D Ground Wellhead Leachate **Base Bottom** Leach **North / South Section GW22** 1399199 4913107 17.4 0.900 18.3 8.750 10.320 9.6 8.0 **GW21** 1399218 4913062 22.3 2.295 24.6 14.220 16.145 10.4 8.5 **GW17** 2.050 8.570 18.825 16.7 1399241 4913005 23.2 25.3 6.4 **GW13** 1399261 4912955 1.440 23.7 6.063 12.072 17.7 22.3 11.7 **GW14** 1399278 4912901 24.0 1.580 25.6 3.415 13.365 22.2 12.2 **GW15** 1399295 4912860 22.8 0.890 23.7 1.850 26,000 21.8 -2.3 **Leach Riser** 1399303 4912843 23.0 2.280 25.3 8.970 9.925 16.3 15.4 GW9 1399349 4912853 23.2 1.378 24.6 10.470 10.470 14.1 14.1 **West / East Section** GW36 22.5 1399161 4912919 1.130 23.6 9.450 11.460 12.2 14.2 **GW37** 1399190 4912876 20.7 1.310 22.0 2.835 5.923 19.2 16.1 **GW26** 1399230 4912911 24.0 1.500 25.5 8.840 10.090 16.7 15.4 **GW11** 1399315 4912949 23.2 0.845 24.0 14.873 26.000 9.2 -2.0 GW7 1399373 4912945 22.5 1.310 23.8 15.000 20.295 8.8 3.5 GW1 1399413 4912962 20.2 0.100 20.3 9.832 10.5 8.7 11.630 Groundwater Wells 1.400 Piez₁ 1399502 4912952 7.8 0.150 8.0 5.500 6.6 2.5 **Gas Compond** 7.1 0.220 7.3 1399363 4913043 1.060 3.035 6.3 4.3 Piez₂ 1399351 4913221 6.9 0.750 7.7 2.205 5.525 5.4 2.1 Piez₃ 1399211 4912667 10.2 0.100 10.3 1.590 8.7 5.5 4.770 Piez4 1399085 4912747 0.300 3.480 7.860 8.2 11.4 11.7 3.8

Gaswell and GW Well Dipping for Leachate & Base Levels - Aug 2022

**GW15 was re-dipped by GHD, the well depth was measured at 26 m bgl*

Figure E.1 Gas Well Locations and leachate measurements

F-1 Introduction

This technical appendix outlines the water balance assessment of rainfall infiltration through the current and proposed cap at Green Island Landfill. The results of this HELP assessment has been incorporated into the SEEP/W modelling (included as Appendix G), which has been prepared to inform an assessment of environmental effects (AEE) of the continued operation and subsequent closure of the landfill.

Rainfall infiltration through the landfill cap was assessed using the Hydrologic Evaluation of Landfill Performance (HELP) software (Berger and Schroeder, 2013). HELP 3.95D is a quasi-two-dimensional hydrologic model for conducting water balance analysis of landfills and cover systems. The model utilises weather, soil and landfill design data to account for the effects of surface storage, runoff, infiltration, evapotranspiration, soil moisture, lateral subsurface drainage, vertical drainage and leakage through soil and liners.

Four different areas of the landfill were modelled to assess leachate generation and leakage. The water balance estimated during the stages of landfill development were used to provide the whole of landfill water balance at different times through the proposed landfill operational lifecycle.

The results of the HELP modelling provide information to support an assessment of environmental effects for continued operation and subsequent closure of Green Island landfill.

F-1.1 Landfill profiles

Four different areas of the landfill, in different phases of landfill development, were considered in predicting the landfill water balance (Figure F.1)

- Area A Transfer station
	- 87.386 m² of final cap
- Area B Operational landfill capped area:
	- 50,379 $m²$ of waste with final cap
	- Area C Perimeter Bund capped area
	- 96,330 m^2 of waste overlain by a capping material
- Area D Operational landfill:
	- Existing conditions through to start of 2022
		- $97,380$ m² of exposed open waste
	- Following installation of mid-landfill drainage layer
		- $97,380$ m² of bare soil contoured towards drainage system
	- Future capped landfill
		- $97,380$ m² of waste with final cap

Figure F.1 Landfill Areas A - D

The profile characteristics adopted are presented in Tables F.8 – F.11 (attached). Soil values of total pore volume, field capacity and wilting point were adopted from published USDA soil textures (Schroeder et al., 1994), however saturated hydraulic conductivity was modified for a number of the soil units. Justification for the adopted soil textures and saturated hydraulic conductivity is presented in Table F.1. Additional scenarios were also run for Area B and Area D where the hydraulic conductivity of the landfill cap (layer 3) was reduced to undertake sensitivity testing (Table F.2).

Table F.1 Justification for adopted soil texture and saturated hydraulic conductivity

1) Schroeder et al., 1994.

2) Berger and Schroeder, 2013.

Table F.2 Scenario testing for landfill cap hydraulic conductivity

F-1.2 Landfill base and drainage layers

HELP models include the functionality to apply the parameters of slope and flow path distance to lateral drainage layers (LDL) that overlay landfill liners or barrier soil layers (BSL) within the landfill profile. The existing Green Island landfill has no formal landfill liner or internal leachate collection system, however the underlying natural geology demonstrates low vertical permeability. Therefore, although a BSL was included within the landfill profiles to simulate the base of the landfill, the overlying layer of municipal waste was modelled as a vertical percolation layer (VPL), not a LDL. Lateral movement of leachate within the waste immediately above the base of the landfill is not expected to be occurring, particularly as the base of the landfill is inferred to be relatively flat.

The design of the operational area of the landfill (Area D) does however include a barrier soil layer and overlying leachate drainage collection system in the middle of the waste profile, which has been installed to capture leachate generated above this layer. The slope and average flow path distance to the leachate collection drains for the modelled LDL is presented in Table F.3. Sensitivity testing of the average flow path distance within the Area D future capped landfill profile indicated that this parameter has no influence on the results as no perching is expected to occur to generate horizontal flow towards the drains. A distance of 15 m was adopted for the Area D future capped profile. For the Area D worst case landfill drainage profile average flow path distances of both 15 m and 75 m were modelled, with the combined results assuming that 1/3 of Area D represents the average conditions in the southern section and 2/3 the average conditions in the northern section.

LDLs are also incorporated above BSLs representing the landfill cap in all landfill profiles which include capping layers (Table F.3). Due to significant differences in the slope and average flow path distance for the eastern and western sections of Area D, two models were used to assess the two different scenarios. The results for Area D (future capped) averaged the results from these two profiles, as each scenario is assumed to represent conditions across approximately 50% of Area D.

Table F.3 Landfill area and slope of drainage layer

F-1.3 Weather and rainfall runoff

Synthetic weather files were generated for a 50-year period using a weather generator model (WGEN), a stochastic model used to generate daily weather variables. Taulis and Milke (2005) from the University of Canterbury developed climate parameters for Dunedin, Otago, which were used within the WGEN to generate inputs for the parameters 'Solar Radiation' and 'Growing Season'. The NIWA weather station located at Musselburgh, Dunedin (Agent numbers 15752 and 5402) was used within WGEN to generate inputs for all other parameters, as described in Table F.4 (NIWA, 2022). Detailed input data are presented in Tables F.15 and F.16 (attached).

Table F.4 WGEN climate input data and sources

In developing the HELP climate data, the user defined parameters and values presented in Table F.5 were also included within the input data for evapotranspiration:

Table F.5 Evapotranspiration input data for WGEN

Rainfall runoff within the HELP model is determined through application of a runoff curve number. The parameters utilised for determination of the curve number include soil texture, surface slope, compaction, surface vegetation and average surface flow path length. The adopted criteria for each modelled landfill profile are presented in Table F.6. As discussed in the previous section, the results for the Area D worst case drainage scenario assume the southern area represents average conditions across 1/3 of Area D and the northern area represents average conditions across 2/3 of Area D. The results for the Area D future capped scenario assume a 50% distribution of the average conditions in the east and west.

Table F.6 Parameters for determination of runoff curve number for each landfill area and development stage

F-2 Results

The results of the HELP modelling are presented in Tables F.17 and F.18 (attached). The reported results represent the annual average output values over the entire model period which incorporates the 50-year stochastic weather files. The results are therefore anticipated to increase or decrease in response to above or below average climatic conditions.

F-2.1 Infiltration

Infiltration as a percentage of total rainfall across the different landfill areas and scenarios are presented in Table F.7.

F-2.2 Consideration of climate change

A review of the estimated changes in precipitation, temperature and relative humidity for Dunedin was undertaken using the NIWA 'Our future climate New Zealand' online tool (NIWA, 2023) and compared against the adopted site climate data from Musselburgh, Dunedin (NIWA, 2022) used to generate the synthetic weather files in WGEN (Section F-1.3). The NIWA (2023) data considered was from the lower end of annual mean change for 2090 (representing end century 2081 – 2100) for the six-model-average RCP8.5. This is considered conservative due to the post-closure consent being due to expire in 2064. The results are presented in Table F.8, Table F.9 and Table F.10. The results indicate approximately a 10% increase in annual mean rainfall between 1995 and 2090, up to 2 degrees C increase in temperature and minimal reductions in relative humidity.

Month	Average monthly rainfall (mm) (1)	Season	Estimated rainfall change $(\%)$ (2)	Estimated monthly rainfall (mm)
January	70	Summer	0%	70
February	61	Summer	0%	61
March	60	Autumn	10%	66
April	57.4	Autumn	10%	63.14

Table F.8 Estimated rainfall change

1) Site climate data from Musselburgh, Dunedin (NIWA, 2022).

2) Our Future Climate New Zealand NIWA (2023). Assumptions for estimated change: RCP8.5, Six-model-average. Adopted change from lower end of annual mean change for 2090 (representing end-century 2081 - 2100).

Table F.9 Estimated temperature change

1) Site climate data from Musselburgh, Dunedin (NIWA, 2022).

2) Our Future Climate New Zealand NIWA (2023). Assumptions for estimated change: RCP8.5, Six-model-average. Adopted change from lower end of annual mean change for 2090 (representing end-century 2081 - 2100).

Table F.10 Estimated temperature change

The 50-year synthetic weather data generated using WGEN (Section F-1.3) is already considered to include climate variability, with the 50-year rainfall data set reporting an average annual precipitation of 754.6 mm, with standard deviation of 93.9 mm. As presented in Figure F.2, annual average rainfall greater than the approximate 10% annual increase estimated by end-century as a result of climate change has already been modelled within the existing results for 12 of the 50 years.

Figure F.2 Stochastic 50-year annual rainfall compared against annual average and 10% increase in annual average.

The greatest rates of leachate generation (through infiltration of rainfall into the landfill waste) are estimated to occur during the Area D 'worst case drainage' scenario. However following capping of Area D the rate of leachate generation is estimated to reduce (Section F-2.1). As capping of the landfill is proposed to be complete by 2029, only the capped landfill will be subject to the estimated changes in climate. The estimated increase in precipitation as a result of climate change is therefore considered likely to result in rates of leachate generation no greater than estimated prior to completion of landfill capping.

F-2.3 References

Berger, K. and Schroeder, P.R. 2013. The Hydrologic Evaluation of Landfill Performance (HELP) Model. User's Guide for HELP-D (Version 3.95D). $6th$ revised edition for version HELP 3.95D.

NIWA, 2022. The national climate database. **https://cliflo.niwa.co.nz/** [Accessed November 2022]

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Schroeder, P. R., Dozier, T.S., Zappi, P. A., McEnroe, B. M., Sjostrom, J.W., and Peyton, R. L. (1994). "The Hydrologic Evaluation of Landfill Performance (HELP) Model: Engineering Documentation for Version 3,"

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* U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

** VPL = Vertical Percolation Layer. BSL = Barrier Soil Layer.

Scenario	Layer No.	Layer Description	Soil Texture Description*	Layer Type**	Layer Thickness (cm)	Saturated Hvdraulic Conductivity (cm/s)	Total Pore Volume (Vol/Vol)	Field Capacity (Vol/Vol)	Wilting Point (Vol/Vol)
Operational		Topsoil	Silty clay - slightly compacted	VPL	10	$2.5E-05$	0.479	0.371	0.251
$landfill -$ 2 capped area 3 $\overline{4}$ 5		Subsoil	Silty clay - slightly compacted	LDL	10	2.5E-05	0.479	0.371	0.251
		Clay cap	Silty clay - moderately compacted	BSL	150	$1.0E-06$	0.452	0.411	0.311
		Soil	Silty clay - slightly compacted	VPL	30	$2.5E-05$	0.479	0.371	0.251
		Municipal waste	Municipal waste	VPL	2000	$1.0E-05$	0.671	0.292	0.077
	6	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311

Table F.12 Area B Operational landfill – existing capped HELP Model soil profile

* U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

** VPL = Vertical Percolation Layer. LDL = Lateral Drainage Layer. BSL = Barrier Soil Layer.

Table F.13 Area C Perimeter Bund HELP Model soil profile

* U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

** VPL = Vertical Percolation Layer. BSL = Barrier Soil Layer.

Scenario	Layer No.	Layer Description	Soil Texture Description*	Layer Type**	Layer Thickness (cm)	Saturated Hydraulic Conductivity (cm/s)	Total Pore Volume (Vol/Vol)	Field Capacity (Vol/Vol)	Wilting Point (Vol/Vol)
Operational $\mathbf{1}$ Daily / landfill- intermediate open waste cover			Sandy loam - slightly compacted	VPL	10	5.0E-04	0.452	0.19	0.085
	$\overline{2}$	Municipal waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
	3	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311
Operational	$\mathbf{1}$	Drainage layer	Gravel	VPL	10	3.0E-01	0.397	0.032	0.013
landfill- worst case	2	Drainage layer	Gravel	LDL	20	3.0E-01	0.397	0.032	0.013
drainage 3 $\overline{4}$ 5		Soil below drainage	Silty clay - Moderately compacted	BSL	500	1.0E-05	0.452	0.411	0.311
		Municipal waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
		Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311
Operational	$\mathbf{1}$	Topsoil	Silty clay - slightly compacted	VPL	15	2.5E-05	0.479	0.371	0.251
landfill - future	\overline{c}	Subsoil	Silty clay - slightly compacted	LDL	20	2.5E-05	0.479	0.371	0.251
capping 3 $\overline{4}$		Clay cap	Silty clay - moderately compacted	BSL	60	1.0E-05	0.452	0.411	0.311
		Intermediate cap	Silty clay - slightly compacted	VPL	30	2.5E-05	0.479	0.371	0.251
	5	Municipal Waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
6 $\overline{7}$		Drainage layer	Municipal waste	LDL	200	1.0E-05	0.671	0.292	0.077
		Soil below drainage	Silty clay - Moderately compacted	BSL	500	1.0E-05	0.452	0.411	0.311
	8	Municipal waste	Municipal waste	VPL	1000	1.0E-05	0.671	0.292	0.077
	9	Natural ground	Silty clay - moderately compacted	BSL	500	5.0E-07	0.452	0.411	0.311

Table F.14 Area D Operational landfill – open waste, worst case drainage and future capped HELP Model soil profiles

* U.S. Department of Agriculture (USDA) soil textural classification system reported in Schroeder et al., (1994).

** VPL = Vertical Percolation Layer. LDL = Lateral Drainage Layer. BSL = Barrier Soil Layer.

Table F.15 Average monthly values used in 50 – year synthetic weather file generation for daily precipitation and daily mean temperature (WGEN)

Table F.16 Parameter values adopted in the solar radiation and evapotranspiration input data files

Table F.17 HELP Model Leachate Results for closed landfill areas

Table F.18 HELP Model leachate results for operational landfill areas

G-1 Groundwater modelling (SEEP/W)

G-1.1 Model set-up

The groundwater assessment included 2D modelling using Geostudio 2021 SEEP/W finite element modelling software. Modelling was undertaken to estimate the seepage into the leachate collection drain and to simulate the leachate head within the landfill.

Two SEEP/W cross-sections were created to model the landfill. The location of the cross-section lines is shown in Figure G.1. The models were created based on a drone survey provided by DCC. Each model was initially run under steady-state conditions to simulate the interpreted baseline groundwater conditions. Model were calibrated to:

- Measured leachate head within the landfill (average level).
- Combined pump station flows (from leachate collection trench), under dry weather conditions.

Steady state model scenarios were then run to simulate future conditions at closure, in particular, capping of the landfill and installation of a leachate collection drain in the southern valley. The Line 2 future scenarios include additional filling as outlined in the design report.

Figure G.1 Location of SEEP/W cross sections

The SEEP/W models adopt the recharge rates as calculated from the HELP modelling assessment (Appendix F), however the recharge rates have been scaled to reflect the actual area and volume of landfill covered by each cross section (Figure G.2 and G.3). The recharge applied in the model reflects to the proportion of rainfall that infiltrates the proposed landfill cap, and has been applied to the layer below the landfill cap (except in un-capped areas where it has been applied to surface). Recharge rates and scaling factors are described in Section G.1.3.

Both model cross sections intercept the existing leachate trench and proposed southern valley trench. The estimated flow rates (per 1 m of trench) into these trenches are presented in the results. These results have been scaled up by the length of the trench.

To augment the leachate interception trench, the current landfill operation has progressively installed trenched leachate drains over the intermediate cover soils in the location of the southern progression of waste placement as well as in the northern sector of waste placed in 2019-2022. It is proposed that additional shallow (0.5 m) leachate drains will be installed in the fill material in the southern portion of the landfill – the area intersected by Line 2. However, given the heterogeneity of the fill materials it is considered that the effect of the drains may be localised. The steady-state model used in this assessment does not simulate variable recharge (from rain events) nor the fill heterogeneity and therefore may overestimate the effect of the drains in reducing leachate head within the fill. For this reason, the future modelling scenarios exclude the shallow leachate drains.

The effect of future sea level rise was considered in the future modelling scenarios. Sea level rise effects were modelled with a higher (0.5 m) river level boundary condition. For this scenario the road/bund level was also increased by 0.5 m as recommended in the *Design Report*.

Figure G.2 Line 1 – SEEP/W model set up (current scenario)

*Figure G.*3 *Line 2 – SEEP/W model set up (current scenario)*

G-1.2 Material Properties

Material properties used in the model scenarios considered are summarised in Table G.1.

Hydraulic conductivity values used in modelling were based on the results of hydraulic conductivity testing undertaken as part of the site investigations. In soil units where test results were not available, widely reported in the literature values were used to approximate hydraulic conductivity values.

It is likely that the waste material is highly heterogeneous, with some areas having a higher hydraulic conductivity than other areas depending on the type of waste materials and the degree of compaction undertaken. Hydraulic conductivities of municipal solid waste (MSW) reported in the literature vary between 1 x 10 \cdot 3 and 1 x 10 \cdot 9 m/s, although most values are in the range of 10⁻⁵ to 10⁻⁶ m/s (LANDSS¹⁰). It is understood that soil makes up a significant proportion of the waste material at GILF, which is likely to reduce the bulk hydraulic conductivity of the landfill material. The best calibration was achieved using a waste hydraulic conductivity of 6 x 10 \cdot 7 m/s.

In general, for natural soils/materials, the hydraulic conductivity perpendicular to the soil layers is slower than the conductivity parallel to soil layers. Therefore, the underlying estuarine and marine sediments (UKEM/LKEM) were assigned a Ky/Kx ratio (anisotropy) of 0.1.

A section of the landfill has recently been capped (2022), however the base case model scenarios do not include the final capping material as the current leachate head level is unlikely to reflect these recent changes.

Table G.1 Material properties

G-1.3 Boundary Conditions

Far field constant head boundaries were set on the landward edge of the model to approximate inferred groundwater level. River level was inferred based on site observations (average water level when the estuary mouth is not blocked) (Table G.2).

¹⁰ Landfill (Aftercare) Decision Support System, University of Southampton **https://landss.soton.ac.uk/**

Table G.2 Boundary conditions used in SEEP/W modelling

*Higher river levels occur from time to time when the river/estuary mouth is blocked

Rainfall recharge was applied to both the landfill and bund surfaces. For the future capped landfill scenarios, the rainfall infiltration was applied to the underside of the landfill cap, as this value represented the rainfall going through the cap (taking into account rainfall and evapotranspiration). The amount of rainfall recharge was based on the outputs of the HELP modelling assessment. However, as the SEEP/W modelling assessment utilises two cross sections to represent a 3D landform, the recharge was scaled down to reflect the average infiltration rate over each landfill sub-area. Without scaling, the models would have overestimated the infiltration into the landfill material and leachate trench flow rates. However, this approach (averaging of rainfall recharge) may underestimate the leachate head level in the centre of the landfill. Scale factors are shown in Table G.3 with scaled infiltration rates included in Table G.4.

טנט שומא <i>Recharge scare ractors</i>								
Area		Area (ha)	Landfill section length (in model, m)	Length of trench covered by model section (m)	Area represented by model (ha)	Recharge scale factor		
Area 1 (Section Line 1)	Landfill	12.9	480	400	19.2	0.67		
	Bund	1.4	50		2.0	0.7		
Area 2 (Section Line 2)	Landfill	11.0	420	620	26.0	0.42		
	Bund	2.5	50		3.1	0.8		

Table G.3 Recharge scale factors

Table G.4 Scaled rainfall recharge applied in SEEP/W models

Section Line	Material / Surface	Scaled infiltration rate (% of annual rainfall)	
Section Line 1	Uncapped	21%	
	Cap (10^{-7} m/s)	21%	
	Cap (10^{-8} m/s)	17.5%	
	Cap (10^{-9} m/s)	3%	
	Bund	14%	
Section Line 2	Uncapped	13%	
	Cap (10^{-7} m/s)	13%	
	Cap (10^{-8} m/s)	12%	
	Cap (10^{-9} m/s)	2.5%	
	Bund	16%	

G-1.4 Model assumptions and limitations

- It is assumed that hydraulic conductivities for the landfill and bund materials adopted in the model simulations are appropriate average values, given the likely heterogeneity of the fill materials.
- The field investigations have identified significant variation in the underlying marine and estuarine sedimentary deposits (UKEM/LKEM). Bulk hydraulic conductivities have been applied to simulate the measured pumping rates, however it is still possible that there may be localised deviation from the simulated results.
- It is assumed that the underlying sedimentary deposits are anisotropic given the nature of the depositional environment.

G-1.5 Model results – existing conditions

Leachate Trench

The results of the modelling for the base case (current) simulations are summarised in Table G.5. The average combined flow rate to the leachate trench is approximately 1-2 L/s, this equates to a rate between 6×10^{-7} m³/s and 1.2 x 10⁻⁶ m³/s per metre of trench (based on a trench length of 1674 m). The results of the base case simulations are within this range, slightly higher flow rates are estimated for Section Line 1, likely due to the proximity to Kaikorai Stream.

As noted in section, 2.2.3, higher flows do occur periodically, associated with wet conditions and additional stormwater flows which are diverted to the leachate system.

Section	Modelled flow rate into trench (m^3/m) of trench)
Line 1	1.4×10^{-6}
Line 2	7.5×10^{-7}

Table G.5 Model results – flow into leachate collection trench

The relative proportion of flows from each side of the trench was calculated from the models (Table G.6). Approximately 70% of the flow to the trench comes from the landfill (and underlying groundwater) and 30% from the direction of the stream. The presence of the HDPE liner in the trench limits flows from surface water, however the model indicates water flow from the stream side into the base of the trench.

Table G.6 Relative proportion of flows into leachate collection trench

The models were used to test the effect on surface and groundwater if the leachate trench was not operational (i.e pumps turned off). As the modelling has been undertaken on a steady state basis, the model assumes that the leachate trench has never been operational, when in reality it is likely to be short term failure only. Therefore, the leachate breakout shown in Figure G.4 is unlikely to be realised with a short term failure. Furthermore, it is considered unlikely that all pump stations would be out of action for an extended period given historical performance and additional mitigation measures that are proposed in the *Design Report*. However, for the purposes of this assessment, the models can be used to estimate the rate of seepage into surface water based on the permeability of the sedimentary units. The model estimates a seepage rate between 4.2 x 10⁻⁷ and 4.7 x 10⁻⁷ $m³/s$ per m of stream, this equates to a rate of \sim 0.5 L/s for the entire stream adjacent to the site (approximately length of 1 km).

Figure G.4 Section Line 1 – leachate and groundwater flow paths with (top) and without trench (bottom) operational.

Leachate Head (in landfill)

The simulated water level (representing landfill leachate) in shown in Figure G.5. The modelled leachate level in the centre of the land fill is 16 m RL in Section Line 1. This is similar to the level recorded in gas wells GW17, 13 and the leach riser (16-17 m RL) (see Appendix D) but less than recorded in GW14 and GW 15 (22 m RL). As noted above, the rainfall recharge rates were scaled to represent the actual area represented by the model section. This approach may result in averaging or flattening of the modelled leachate head. However, in general the models predict leachate close to the surface in the bund materials. Observations made during a test pit

investigation of the bund materials (Tonkin and Taylor, 2020) suggest that the leachate level is close to the surface in some areas. It is understood that leachate breakout from the bund has occurred from time to time, but has been addressed by the installation of gravel drains in these areas.

In Line 2, the simulated leachate level in the centre of the section is approximately 10 m RL. As the leachate survey focussed on older parts of the landfill, there is limited information available for calibration in this area. The leachate level recorded in Piezo 3 and Piezo 4, located south of the section line, was 8.7 and 8.2 m RL respectively.

Figure G.5 Leachate head in landfill -Line 1 (top) and Line 2 (bottom) – current scenario

G-1.6 Future modelling scenarios

The base case SEEP/W models were modified to simulate conditions at closure (capping and leachate management). The future scenarios tested three different cap permeabilities (10⁻⁷ to 10⁻⁹ m/s). Additional scenarios were run to test the effect of trench failure and sea level rise effects. The model scenarios are listed in Table G.7. The future modelling scenarios include the installation of the proposed southern valley trench.

Scenario Number	Hydraulic conductivity of cap		Description
	Currently uncapped*	Capped areas (north/central, applies- Section Line 1 only)	
1A	1×10^{-7} m/s	1×10^{-8} m/s	Leachate trenches operational
1B	1×10^{-7} m/s	1×10^{-8} m/s	Trench" turned off"
1 ^C	1×10^{-7} m/s	1×10^{-8} m/s	Trench operational, Sea level rise (+ 0.5 m) applied to river boundary.
2A	1×10^{-8} m/s (all areas)		Leachate trenches operational
2B	1×10^{-8} m/s (all areas)		Trench" turned off"
2C	1×10^{-8} m/s (all areas)		Trench operational, Sea level rise (+ 0.5 m) applied to river boundary.
3A	1×10^{-9} m/s (all areas)		Leachate trenches operational

Table G.7 Future modelling scenarios

*Only applies to areas not currently capped. It is understood that the northern and central part of the landfill (Zone B in Figure G.1, Appendix G) has recently been capped with low permeability materials (achieving 10⁻⁸ 10⁻⁹ m/s in lab tests). For the purpose of this assessment, it has been assumed that the hydraulic conductivity is 10 8 of this new cap is at least 10-8 m/s.

Current and future scenario modelling results are summarised in Tables G.8 and G.9. Model flow rates (per metre) are scaled up by the length of trench each model section covers. This does not include the leachate trench flows from pump stations 5 to 8 (transfer station area) which contribute to the overall dry weather flow rate of 1-2 L/s. However, trench flow rates are expected to be lower than in the modelled sections, as the volume (and height) of filling is less.

Surface water interaction

The models were used to estimate the flow rate into or from (stream depletion) Kaikorai Stream. These results are shown Table G.9. In Section Line 1, the interaction with the stream is reversed depending on whether the leachate drain is "turned on".

For Section Line 2, the model predicts flows from groundwater into the stream in all model scenarios. The flow rate is more than an order of magnitude higher when the drain is turned off than when it is operating, showing that the leachate trench is effective in intercepting leachate/groundwater from the landfill. When the leachate trench is operating, flow vectors indicate the groundwater flow into the stream $(-2 \times 10^{-8} \text{ m}^3/\text{s}$ per m, 0.01 L/s for a 600 m length) represents upward groundwater seepage from the underlying estuarine sediments. The stream is located ~70 m from the trench, therefore while the trench appears be effective at intercepting leachate from the landfill, it does not result in stream depletion and does not influence groundwater levels immediately adjacent to the stream.

Table G.8 SEEP/W model results – leachate trench and leachate head

Flow rate (per metre) multiplied by length of trench represented by each section:

Line 1 – Perimeter trench, 400 m, Southern Valley trench (150 m)

Line 2 – Perimeter trench, 620 m, Southern Valley trench (300 m)

Table G.9 SEEP/W model results – leachate trench and leachate head

Flow rate (per metre) multiplied by length of stream/trench represented by each section:

Line 1 – Perimeter trench, 400 m

Line 2 – Perimeter trench, 620 m

G-1.7 SEEP/W Modelling Summary

A summary of the key finding from the modelling assessment is provided below in Table G.10.

Table G.10 Modelling summary

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